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Abstract

Pharmaceuticals and analogs of bisphenol A (BPA) are increasingly threatening environmental pollutants. In this study, mixtures of selected pharmaceuticals (diclofenac sodium salt, chloramphenicol, oxytetracycline hydrochloride, fluoxetine hydrochloride, estrone, ketoprofen, progesterone, gemfibrozil and androstenedione) were prepared with BPA and its two analogs (namely, bisphenols F and S) at such ratios to reflect environmentally detectable levels. Then, the mixture solutions were studied with a XenoScreen YES/YAS assay to determine the variations in the initial hormonal response of each pharmaceutical compound due to the presence of a bisphenol analog. The results obtained were modeled with the concentration addition (CA) and independent action (IA) approaches, the trueness of which was studied with model deviation ratios (MDR). The estrogenic agonistic activity of the drugs studied was most strongly affected by the presence of BPA in solution (twenty-one cases of synergy observed for CA models versus twelve cases of antagonism in the case of IA predictions). BPS shows a strong agonistic estrogenic impact on most of the drugs studied at medium and high concentration levels; androgenic agonistic activity was also impaired with elevated concentrations of BPS. Increasing the concentration of BPF in a reaction mixture also increased the number of YES+ synergism incidences (for CA modeling). Estrone, progesterone and androstenedione were mostly affected by the highest BPF concentrations studied in the case of androgenic agonistic research performed.

Keywords	bisphenol A, bisphenol A analogues, pharmaceuticals, endocrine disruptors of mixtures, environmental pollution
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Bisphenols (A, S, and F) affect the basic hormonal activity determined for pharmaceuticals – study of Saccharomyces cerevisiae

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8 Abstract: Pharmaceuticals and analogs of bisphenol A (BPA) are increasingly threatening 9 environmental pollutants. In this study, mixtures of selected pharmaceuticals (diclofenac sodium salt, chloramphenicol, oxytetracycline hydrochloride, fluoxetine hydrochloride, estrone, ketoprofen, 10 11 progesterone, gemfibrozil and androstenedione) were prepared with BPA and its two analogs 12 (namely, bisphenols F and S) at such ratios to reflect environmentally detectable levels. Then, the 13 mixture solutions were studied with a XenoScreen YES/YAS assay to determine the variations in the 14 initial hormonal response of each pharmaceutical compound due to the presence of a bisphenol 15 analog. The results obtained were modeled with the concentration addition (CA) and independent 16 action (IA) approaches, the trueness of which was studied with model deviation ratios (MDR). The 17 estrogenic agonistic activity of the drugs studied was most strongly affected by the presence of BPA 18 in solution (twenty-one cases of synergy observed for CA models versus twelve cases of antagonism 19 in the case of IA predictions). BPS shows a strong agonistic estrogenic impact on most of the drugs 20 studied at medium and high concentration levels; androgenic agonistic activity was also impaired with elevated concentrations of BPS. Increasing the concentration of BPF in a reaction mixture also 21 22 increased the number of YES+ synergism incidences (for CA modeling). Estrone, progesterone and 23 androstenedione were mostly affected by the highest BPF concentrations studied in the case of 24 androgenic agonistic research performed.

*Keywords*: bisphenol A, bisphenol A analogs, pharmaceuticals, endocrine disruptors of mixtures,
 environmental pollution

*Capsule*: Analogs of BPA affect endocrine potential of numerous pharmaceuticals; the extent andmagnitude of disruption is compound and concentration dependent.

1. Introduction

Comprehensive studies on the environmental fate and impact of pharmaceuticals (and their residues) as well as plasticizers on living organisms are a relatively new research topic. Unfortunately, many studies and literature reports in this area concern only the instrumental quantification of the

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presence of a given xenobiotics (and in best case scenario - possible transformation products). Certainly, this type of information is very useful, but may be insufficient to determine the impact of given pharmaceuticals on living organisms existing in a particular environment. Most importantly, residues of pharmaceuticals occur in the environment in mixtures with other pollutants, and the degree of their toxicity may vary depending on the influence of environmental factors (such as pH, salinity, and ion co-presence (Wieczerzak et al. 2018).

39 Bisphenols (BPs) are quite widely used and have been detected in significant amounts in 40 drinking water, foods (especially pre-packaged foods) and beverages, as well as in environmental samples (Eladak et al. 2015; Danzl et al. 2009). In 2015, Corrales et al. have reviewed over 500 41 42 publications dealing with the prevalence of BPA, noting that the average amount of BPA detected in 43 wastewater treatment plants (WWTP) effluents worldwide ranged from 5 up to 370  $\mu$ g/L, and 44 surface waters samples may contain even 56 µgL<sup>-1</sup> of BPA; however, salty waters of the Baltic Sea 45 may contain up to 67.7 ng/L according to Staniszewska et al. (2014). Lower levels of BPA were reported in potable tap water samples levels, up to  $1.3 \,\mu$ g/L (France) (Colin et al. 2014). The amount 46 47 of BPA polluting sludges and biosolids has been reported to vary from 10 to > 100 000  $\mu$ g/kg in 48 industrial areas (dry weight) (Meesters and Schroder 2002).

49 The literature is not rich in data on BPA analogs; the analysis of wastewater from WWTPs in 50 Slovakia revealed that only BPF and BPS (but not other BPs) were detected in the WWTP influents (in 51 amounts of 36.7 ng/L and 40.6 ng/L, respectively) (Česen et al. 2018). Another study detected 52 concentrations of BPF as high as 2850 ng/L in the Tamagawa River in Japan (Yamazaki et al. 2015). 53 Liao and Kannan in 2013 studied a variety of food samples from the US and determined that the 54 highest amount of BPF (1130 ng/g) was present in samples of mustard (dressing) and ginger; 55 additionally, the studies confirmed that canned foods contained higher concentrations of bisphenols 56 than foods stored in glass, paper or even plastic containers (Liao and Kannan 2013).

57 The largest load of pharmaceuticals is introduced into the environment as a result of 58 activities of the pharmaceutical industry, hospital facilities and households (Santos et al. 2010), 59 veterinary and agriculture as well as in animal breeding (Li 2014). Analgetics, anti-inflammatory drugs 60 and anticoagulants are the most frequently detected groups of pharmaceutical residues; ibuprofen, diclofenac and ketoprofen (active ingredients of non-steroid anti-inflammatory drugs) are most 61 62 frequently found in soil and water samples. The second most frequently detected group of 63 pharmaceutical residues comprises antibiotics and antifungal agents such as oxytetracycline and 64 chloramphenicol, which have been detected in surface waters of Nigeria at concentration levels of 65 460 ng/L and 600 ng/L, respectively (Olatunde et al. 2014). Androgens as well as estrogens such as androstenedione, progesterone and estrone have been detected in surface and wastewaters in the 66 67 following amounts: >100 ng/L, 66 ng/L, and 36 ng/L, respectively (Kim et al. 2007; Chang et al. 2011). 68 In environmental samples, residues of anti-epileptic and antidepressant drugs have also been found; 69 among them, fluoxetine is one of the most frequently detected, and its average surface water 70 concentrations ranges globally from 0.012 to 1.4  $\mu$ g/L (Weinberger and Klaper 2014). Studies have 71 shown that lipid regulators (such as gemfibrozil) are also present in the environment at 72 concentration of 70.27 ng/L e.g. in the Llobregat River (Osorio et al. 2016).

73 Environmental contamination still seems to be underestimated by policy makers and 74 entrepreneurs. Scientists continue to learn how to better quantitatively and qualitatively determine 75 the concentrations of pollutants in complex mixtures to assess acute and chronic exposure of 76 different organs/tissues/organisms to single compounds belonging to different groups and to predict 77 their impact on ecosystems. Tanaka et al. (2018) confirmed that a time-weighted average 78 concentration exposure to pyriproxyfen during a period of sensitivity affects the sex ratio, causing 79 approximately equivalent population-level effects as reproductive inhibition regardless of the 80 exposure scenario. Wan et al. (2018) concluded that one in-unit increase in urinary BPS was 81 correlated with a 0.72-day increase in pregnancy duration, and in case of fetal sex each in-unit 82 increase in maternal urinary BPS was associated with increased gestational age; on the other hand, in 83 this study no associations of BPS with birth weight or length were found. Chen et al. (2018) 84 confirmed the presence of seven bisphenols and TCS in 283 urine samples collected from children 85 from South China aged between 3 and 11 years and noticed that age, but not gender, was negatively 86 associated with urinary levels of BPA and BPS.

87 Synergism and antagonism are two basic phenomena that may describe effect of one 88 substance/factor combination (Wieczerzak et al. 2015 and 2016). Among several mathematical 89 models enabling prediction if any of the two phenomena occurred the concentration addition (CA) 90 and independent action (IA) are most commonly used. Briefly the CA model is used to study if analytes in a mixture exhibit a similar mode of action. The concept is that the similarly acting 91 92 substances act jointly in an additive manner when present together in mixture. The IA model should be used to test toxicants of dissimilar mode of action when present in a mixture - assumption is done 93 94 that they act independently. This model is rather a form of a statistical approach to predict if one of 95 multiple events could occur.

96 In this study, BPS and BPF were selected as plausible mixture components (next to BPA) due 97 to fact that manufacturers are gradually replacing BPA with these compounds to comply with restrictions and regulations on BPA (Chen et al. 2016, IRIS 1988, Rochester 2013). BPS is also used as an anticorrosive agent in epoxy glues or as a reagent in polymer reactions; the presence of BPS has also been confirmed in canned foodstuffs. BPS is less prone to environmental degradation than BPA and BPF and has been found in over 80% of the urine samples studied. BPF, on the other hand, is used in manufacturing to increase the durability and thickness of the materials produced (not to

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103 mention its usage in liners, lacquers, adhesives, plastics, coating of drink containers and food cans 104 and dental sealant). BPF was confirmed to cause genotoxicity when introduced to Hep G2 cells, and 105 its endocrine potential is also well confirmed. Due to the confirmed endocrine potential of numerous 106 pharmaceuticals (Wieczerzak et al. 2016) and BPA analogs (Owczarek et al. 2018a and 2018b), it is 107 important to study the behavior of these common environmental pollutants when present in 108 mixtures, which was the goal of this study.

109 2. Materials and methods

110 2.1. Chemicals and reagents

111 The model substances selected for the study, diclofenac sodium salt (diclofenac s.) (CAS no. 153907-79-6), chloramphenicol (CAS no. 56-75-7), oxytetracycline hydrochloride (oxytetracycline h.) 112 113 (CAS no. 2058-46-0), fluoxetine hydrochloride (fluoxetine h.) (CAS no. 56296-78-7), estrone (CAS no. 114 53-16-7), ketoprofen (CAS no. 22071-15-4), progesterone (CAS no. 57-83-0), gemfibrozil (CAS no. 115 25812-30-0), androstenedione (CAS no. 63-05-8), bisphenol A (CAS no. 80-05-7), bisphenol S (CAS no. 116 80-09-1) and bisphenol F (CAS no. 620-92-8), were purchased from Sigma Aldrich (Germany), and 117 they were of analytical grade (>99%). A set of XenoScreen YES/YAS reagents was purchased from 118 Xenometrix AG (Switzerland), namely, a vial with hER $\alpha$  (YES) yeasts (to determine estrogenic activity) 119 and hARa (YAS) yeasts (to determine androgenic activity) immobilized and dormant on the filtration 120 paper, basal medium, vitamin solution, L-aspartic acid solution, L-threonine solution, CuSO<sub>4</sub> solution, 17β-oestradiol (E2, YES+ (agonist) control), 5α-dihydrotestosterone (DHT, YAS+ (agonist) control), 4-121 122 hydroksytamoxyphene (HT, YES- (antagonist) control), and flutamide (FL, YAS- (antagonist) control). 123 DMSO (dimethyl sulphoxide, CAS no. 67-68-5) and CPRG (chlorophenol red- $\beta$ -D-galactopyranoside) 124 (CAS no. 99792-79-9) dye were purchased from Sigma Aldrich (Germany). Measurements of the OD<sub>690</sub> cell density (wavelength 690 nm) and of the intensity of the CPRG transformation product 125 126 OD<sub>570</sub> (wavelength 570 nm) were performed with a TECAN Infinite M200 spectrophotometer.

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## 2.2. XenoScreen YES/YAS methodology

To investigate the endocrine potential of the solutions, a slightly modified protocol for 130 XenoScreen YES/YAS was utilized. Saccharomyces cerevisiae yeast cells were cultivated from the filter papers in a growth medium (basic medium with a vitamin solution and a solution of L-threonine, L-131 aspartic acid and copper (II) sulfate (VI)). Then, 5 mL of the growth medium was transferred to 132 133 labeled culture bottles with caps with a gas permeable filter. Afterwards, the yeast disks were 134 sterilely transferred and placed on an orbital shaker set at 32 °C and 100 rpm for 48 h. Then, 100 µL of DMSO was added to each control vial containing the standards: E2 ( $17\beta$ -oestradiol control of YES agonist), DHT (5a-dihydrotestosterone control of YAS agonist), HT (4-hydroxytamoxifen control of

137 YES antagonist), and FL (flutamide control of YAS antagonist). Test plates were prepared in such a 138 way that the controls were in duplicate in eight serial dilutions:

- YES agonist plate E2 (min. concentration 1.10<sup>-11</sup> M, max. concentration 1.10<sup>-8</sup> M), 139

- YES antagonist plate HT (min. concentration 1.10<sup>-8</sup> M, max. concentration 1.10<sup>-5</sup> M; additionally, in 140 141 the entire plate, E2 was present at a constant concentration of  $1 \cdot 10^{-9}$  M),

- YAS agonist plate DHT (min. concentration 1.10<sup>-9</sup> M, max. concentration 1.10<sup>-6</sup> M), 142

143 - YAS antagonist plate FL (min. concentration 1x10<sup>-7</sup> M, max. concentration 1·10<sup>-4</sup> M; additionally, in 144 the entire plate, DHT was present at a constant concentration of  $3 \cdot 10^{-8}$  M).

145 The addition of E2 or DHT at the same concentration as the entire YES or YAS antagonist 146 plate, respectively, was intended to help examine (confirm/deny) the andro- and estrogenic 147 antagonistic activity of the samples. A substance with antagonist properties competes with the E2 or 148 DHT present on the plate and binds to the receptor without inducing the expression of  $\beta$ -149 galactosidase. Without the enzyme, substrate staining does not occur. However, if the test sample 150 does not contain antagonistic substances, then the E2 and DHT present in the wells bind with the 151 receptor, expressing  $\beta$ -galactosidase; thus, the staining of the substrate occurs.

152 Twenty microliters of adjusted drug sample and sixty  $\mu$ L of 6 mM CRPG dye were added to 153 each assay well. Pharmaceuticals were mixed in three concentration ratios in such a way as to detect 154 a broad range of possible interactions. All of the studies on the mixtures were performed in 155 triplicate; furthermore, controls were made for pure substances in duplicate, and YES and YAS 156 suspensions of yeast cultures (100  $\mu$ L; yeast cell density >0.3 OD<sub>690</sub>) were added into the agonist and 157 antagonist YES and YAS plates, respectively. Assay plates were sealed with semipermeable 158 membranes and placed in the zipper bag moistened with watered gauze on an orbital shaker for 48 h 159 at 32 °C and 100 rpm. After 48 h of incubation, the cell density determined by OD was read at a wavelength of 690 nm, and the color intensity at a wavelength of 570 nm was determined. 160 161 Afterwards, the activity of  $\beta$ -galactosidase was calculated as the ratio of [(OD<sub>570</sub>-OD<sub>690</sub>)/OD<sub>690</sub>]. All 162 experiments were run and measured in triplicate.

163 To determine whether the addition of selected bisphenol analogs to the pharmaceutical 164 solutions would affect the endocrine potential, preconcentrated solutions of pharmaceuticals were prepared. The study of the effects of BPA analogues on the toxicity of the pharmaceuticals was 165 166 conducted at three concentration levels (listed for each substance in Table 1).

2.3. Calculation of MDR values

168 In order to determine whether the presence of one compound in a binary mixture with another substance would affect the endocrine potential against Saccharomyces cerevisiae, the mixtures were prepared such that the compounds were present in appropriate ratios to reflect, respectively, the C1 concentration of first substance with the C1, C2 or C3 concentration of the second substance;

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subsequently the C2 concentration of the first substance with the C1, C2 or C3 concentration of second substance, *etc.* C2 represents the concentration of analyte's solutions calculated from the XenoScreen YES/YAS test, and it is summarized in Table 1. C1 represents the concentrations of substance reduced to 50% of C2, and respectively C3 is the concentration elevated to 150% of C2 – the same scheme was applied to each pharmaceutical studied.

The toxicological effect of a mixture of pharmaceuticals with BPA and its analogues on *S*. *cerevisiae* cells was mathematically assessed with both the Concentration Addition - CA and Independent Action - IA models using equations 1 and 2, respectively (Kudłak et al. 2016, Wieczerzak et al. 2015):

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$$EC_{X_{Mix}} = \left(\sum_{i=1}^{n} \frac{P_i}{EC_{x_i}}\right)^{-1}$$
 (eq. 1.)

# 183 where:

- *ECx<sub>mix</sub>* is the *x<sub>mix</sub>* effect caused by the total concentration of the mixture of studied chemicals
   (components) (Expected value);
- 186  $p_i$  indicates the fraction of component *i* in the mixture, calculated on the basis of the 187 concentration of component *i* in the mixture;
- 188 *n* indicates the number of components in the mixture;
- and *ECx<sub>i</sub>* indicates the *x<sub>i</sub>* effect caused by component *i* at a given studied concentration in
  the mixture.

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$$E(C_{mix}) = 1 - \prod_{i=1}^{n} (1 - E(c_1))$$
 (eq. 2.)

193 where:

- *EC<sub>mix</sub>* is the overall effect expressed as a fraction of the maximal possible effect of a mixture of chemical *i* (expected value);

- *c<sub>i</sub>* indicates the concentration of component *i* in the mixture;

- *n* indicates the number of components in the mixture;
- *E*(*c<sub>i</sub>*) indicates the effect of component *i*, applied separately.

To verify the difference between the predicted and observed effect, the Model Deviation Ratio (MDR) approach was applied, defined as shown in equation 3 (Wieczerzak *et al.* 2016; Kudłak *et al.* 2016):

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$$MDR = \frac{Expected}{Observed} \qquad (eq. 3.)$$

204 where:

*Expected* is the effective toxicity (raw values of genotoxicity calculated according to eq. 1. or 2.) of
 the mixture predicted by the CA or IA model, and

207 - Observed is the effective toxicity (raw values of genotoxicity calculated) for the mixture obtained
208 during the toxicity studies.

MDR values lower than 0.50 are indicators of synergic behavior of substances present in mixture while values >2.00 justify the statement of antagonistic action of analytes studied; MDR values within 0.50-0.71 and 1.40-2.00 mean, respectively, under- and overestimation of calculated models.

213 3. Results and discussion

3.1. Results of hormonal activity studies of selected analytes

Based on previous experience and in relation to published data, the proper concentrations (3 constant levels for each analyte) were selected for the studies (as presented in Table 1) to reflect the environmental concentration levels of the analytes. Then, mixtures of BPA analogs with respective pharmaceuticals were prepared, and their endocrine potential was studied as described in subchapter 2.3.

#### Table 1.

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3.2. Results for BPA impact on the hormonal activity of pharmaceuticals

BPA is a well-known to bind both nuclear estrogen receptors (ER $\alpha$  and ER $\beta$ ) and strongly binds to estrogen related receptors  $\gamma$  (ERR- $\gamma$ ). In other studies it has been shown that BPA also acts as a weak antagonist of androgen receptors (AR) (Stossi et al. 2014). This type of multiple mechanisms of interaction on receptors contributes to the observation of many different effects, both estrogenic and androgenic.

The MDR values presenting a variation of modeled values from measured values in the case of studies of BPA impact on the endocrine potential of selected drugs are given in Table 2 below.

## Table 2.

## 3.2.1. YES+

The estrogenic agonistic activity of the drugs studied seems to be most strongly affected by the presence of BPA in solution. In twenty one cases (mixtures with diclofenac, chloramphenicol, oxytetracycline and fluoxetine) synergy was observed for the CA models, whereas antagonism was observed only for progesterone and gemfibrozil in the case of twelve IA predictions (please refer to
Table C in the electronic supplementary material for details on the percentile values for the MDRs of
the experiments on BPA YES+).

Antagonism was observed for IA model for mixtures of gemfibrozil (at all concentrations levels) with BPA (only for the lowest concentration - C1). Cases of antagonism were observed also for all mixtures of BPA and progesterone. Although the tests performed do not answer the question what is happening at the cellular level, so reference data were needed to better interpret the test results. The literature is a bit skimpy regarding the impact of BPA on gemfibrozil, but it can contain data on the effect of BPA, in clinical trials, on the development of progesterone resistance, which may explain the strong antagonism observed in these mixtures (Aldad et al. 2011).

247 In twenty one cases (mixtures with diclofenac, chloramphenicol, oxytetracycline and 248 fluoxetine) synergy was observed for the CA models with the strongest synergism shown for BPA and 249 oxytetracycline h. pair (please refer to Table C in the electronic supplementary material for details on 250 the percentile values for the MDRs of the experiments on BPA YES+). In these cases the metabolic 251 pathway of these substances may be responsible, because the enzymes from the P450 group (CYP1A, 252 CYP2A, etc.) are involved in the detoxification of all these substances, the occurrence of two 253 substances that involve the same detoxification enzymes may burden the system leading to greater 254 toxicity (endocrine disruption) (Hanioka et al. 2000; DrugBank 2018a, 2018b, 2018c, ). Similar results 255 for agonist interactions were obtained for the combination of these drugs with BPS and BPF. 256 However, this should be verified in further studies.

# 3.2.2. YES-

258 Estrogenic antagonistic activity was not affected by the presence of BPA in a synergic 259 manner; no such cases are observed. Seventeen cases of antagonistic behavior were confirmed for 260 mixtures with chloramphenicol, fluoxetine, progesterone and gemfibrozil with IA model (please refer 261 to Table D in the electronic supplementary material for details on the percentile values for the MDRs of the experiments on BPA YES-). The most antagonistically potent to estrogen receptors was the 262 263 mixture of BPA and fluoxetine (at all concentrations studied); however, the observed phenomenon 264 had a weak character and was strengthened with increasing concentration of both substances 265 (please refer to Table D in the electronic supplementary material for details).

## 3.2.3. YAS+

BPA at the lowest concentration level studied (C1) was adequately modeled with the CA approach (one case of synergism for androstenedione can be noticed, the number of observed cases of synergy increases with increasing concentration of BPA in the mixture) while IA shows several cases of antagonism (for diclofenac, chloramphenicol, oxytetracycline, fluoxetine and gemfibrozil). At higher concentration levels synergism was observable for mixtures with estrone, progesterone and

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androstenedione (with CA approach) showing the importance of BPA in cases of co-presence with these hormonal chemicals. In the case of androgenic agonistic effects, all observed cases of synergy were in mixtures of BPA with hormonal substances. The IA models showed twenty cases of antagonistic action (as presented in table 2) (please refer also to Table E in the electronic supplementary material for details on the percentile values for the MDRs of the experiments on BPA YAS+).

278 3.2.4. YAS-

In the case of androgenic antagonistic studies, no cases of synergism were observed for either the IA or CA models. Interestingly, antagonistic behavior was observed only for the highest concentration level of fluoxetine (four cases) (please refer to Table F in the electronic supplementary material for details on the percentile values for the MDRs of the experiments on BPA YAS-).

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3.3. Results of BPS impact on the hormonal activity of pharmaceuticals

BPS (just like BPA) is able to easily bind to nuclear receptors (the data shows that this bond is weaker than of BPA and estradiol); there also some studies that show androgenic activity of BPS (Rochester and Bolden 2015). The results obtained for BPS and BPF (described below) are similar to what could indicate a similar mechanism of action in mixtures. The MDR values presenting a variation of modeled values from measured ones in the case of studies of BPS impact on the endocrine potential of selected drugs are given in table A (placed in electronic supplementary material).

3.3.1.YES+

292 BPS showed strong agonistic estrogenic impact on most of the drugs studied at a medium 293 (C2) and high (C3) concentration of BPS. Forty cases of synergism (with diclofenac, chloramphenicol, 294 oxytetracycline, fluoxetine, ketoprofen, progesterone and gemfibrozil) and nine cases of 295 underestimation were observed for the CA models. The strongest synergistic effect was observed for 296 gemfibrozil at the level C2 and BPS at the highest concentration (C3). With the increase in BPS 297 concentration, one can observe a "transition" from antagonistic interactions to synergistic properties 298 for progesterone, androstenedione, chloramphenicol and gemfibrozil couples, which would mean 299 that the higher concentration of BPS in these mixtures affect dominantly estrogen receptors, which 300 may be related to the metabolic pathway of these substances. At the lowest concentration level of 301 BPS studied almost no impact on other analytes was found with the possible exception of 302 androstenedione where antagonistic behavior was observed for both models used; with increasing 303 concentration of BPS there was a tendency for underestimating the impact of BPS and even 304 exhibiting synergic action (please refer to Table G in the electronic supplementary material for details on the percentile values for the MDRs of the experiments on BPS YES+). 305

3.3.2. YES-

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The impact of BPS on the estrogenic antagonistic properties of selected pharmaceuticals was adequately modeled with both the CA and IA approaches. IA showed twenty-four cases of antagonistic activity (mostly at the lowest concentration level of BPS) for mixtures with diclofenac, chloramphenicol, progesterone, gemfibrozil and the lowest concentration levels of oxytetracycline and fluoxetine (with the highest MDR values for 2.74 for BPS (C1) and gemfibrozil (C3)); (refer to Tables A and G in the electronic supplementary material for details on the results and on the percentile values for the MDRs of the experiments on BPS YES-).

314 3.3.3. YAS+

Synergism and underestimation were observed in forty five cases of CA modeling while the IA models showed five cases of antagonistic behavior of BPS on drugs when androgenic agonistic behavior is concerned. BPS had a synergic impact on estrone, progesterone and androstenedione, confirming the environmental threats posed by this substance (please refer to Table I in the electronic supplementary material for details on the percentile values for the MDRs of the experiments o BPS YAS+).

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# 3.3.4. YAS-

The androgenic antagonistic impact of BPS was noted in only four cases of IA modeling (for fluoxetine and estrone). In other cases both models adequately forecast the behavior of BPS-drug mixtures (MDR values close to 1.00) (please refer to Table J in the electronic supplementary material for details on the percentile values for the MDRs of the experiments on BPS YAS-).

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## 3.4. Results of BPF impact on the hormonal activity of pharmaceuticals

328 BPF (similarly to its analogs) shows estrogenic and antiandrogenic properties ((Rochester and 329 Bolden 2015). The results obtained were similar to the BPF, both agonistically and antagonistically to 330 estrogen and androgenic receptors, suggesting similarity in MOD. The MDR values presenting a 331 variation of the modeled values from measured values in the case of studies of BPS impact on the 332 endocrine potential of selected drugs are given in table B (in electronic supplementary material).

3.4.1.YES+

The strongest synergistic effect was observed for gemfibrozil at the level C3 and BPF at the highest concentration (C3). As in the case of BPS, with increase of BPF concentration, a "transition" from antagonistic interactions to synergistic properties can be observed for progesterone, androstenedione, chloramphenicol and gemfibrozil pairs. At the lowest concentration level studied BPF did not show any synergic impact on the estrogenic agonistic activity of the drugs selected. Such behavior (for CA modeling) was observable in numerous cases with increasing concentration levels of BPF. Only the activity of estrone was not greatly impacted by the presence of BPF. No antagonistic

behavior of BPF with the CA modeling approach was observed, and seven cases of antagonism were
detected with IA modeling (for progesterone, androstenedione and chloramphenicol) (please refer to
Table I in the electronic supplementary material for details on the percentile values for the MDRs of
the experiments on BPF YES+).

346 3.4.2. YES-

For the estrogenic antagonistic impact of BPF only one case of synergism (with ketoprofen) was observed while sixteen results confirm underestimation by CA modeling. Antagonistic behavior is more likely and was observed in fourteen cases, and forty-four overestimations were noted as a result of IA modeling (refer to table B for details) (please refer also to Table L in the electronic supplementary material for details on the percentile values for the MDRs of the experiments on BPF YES-).

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3.4.3.YAS+

355 Androgenic agonistic endpoints were in most cases not affected by the presence of mixtures 356 of the lowest BPF concentration with the drugs studied. Several exceptions include ten synergic 357 mixtures of BPF with estrone, progesterone and androstenedione and two antagonistic mixtures with 358 diclofenac and chloramphenicol (in IA studies); at higher concentration levels BPF showed a synergic 359 impact on the drugs studied. Thirty nine cases of underestimation were present in the CA studies, 360 and twenty-six cases of overestimation were found for IA modeling (please refer to Table M in the 361 electronic supplementary material for details on the percentile values for the MDRs of the 362 experiments on BPF YAS+).

3.4.4.YAS-

Androgenic antagonistic activity seemed to be only slightly impacted by mixtures of BPF and the drugs studied. Such synergic impact was seen for six mixtures with chloramphenicol while five antagonistic incidences were observed for diclofenac and fluoxetine (for IA modeling) (please refer to Table N in the electronic supplementary material for details on the percentile values for the MDRs of the experiments on BPF YAS-).

369 4. Conclusions

As stated in this research, the variations in endocrine potential are to a great extent concentration dependent. YES+ and YAS+ activity in most cases was affected in a synergic manner while YES- and YAS- were affected in an antagonistic manner in most of experiments performed. As can be easily noticed, the magnitudes of interactions were dependent on the chemicals (BPA, BPS or BPF) admixed in solution as well as their concentration levels. 375 Certainly, the work does not give the ultimate answer on possible interactions of BPA 376 analogues and pharmaceuticals but constitutes one of next steps and guidelines proving the 377 necessity of taking into account phenomena of antagonism or synergism when studying target-378 oriented processes occurring at levels of ecosystems. Future research should focus on studying 379 higher order mixtures of pollutants of persistent character belonging to other chemical groups (e.g., 380 derivatives of bisphenol A, diglycidyl ether, other pharmaceuticals, pesticides, etc.) as well as their 381 degradation products and on wider sets of bioassays with varying biotest endpoints. BPS and BPF are 382 unfortunately considered to be safe replacements of BPA; therefore, their amount in the 383 environment starts increasing and for this reason it is important to monitor them in terms of their 384 properties also in mixtures - what always happens in case of environmental samples (especially due 385 to the fact that the literature data is less rich in information about BPS, BPS than in BPA). In this way 386 the information coming from instrumental studies will be broadened with newly developed, accurate 387 mathematical models describing the environmental threat posed by mixtures of pollutants.

388

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392 References

Aldad, T. S., Rahmani, N., Leranth, C., & Taylor, H. S. (2011). Bisphenol-A exposure alters endometrial
progesterone receptor expression in the nonhuman primate. *Fertility and sterility*, *96*(1), 175-179.

Caban, M., Lis, E., Kumirska, J., Stepnowski, P. (2015). Determination of pharmaceutical residues in
drinking water in Poland using a new SPE-GC-MS (SIM) method based on Speedisk extraction disks
and DIMETRIS derivatization. Science of the Total Environment, 538, 402-411.

Česen, M., Heath, D., Krivec, M., Košmrlj, J., Kosjek, T., Heath, E. (2018). Seasonal and spatial
variations in the occurrence, mass loadings and removal of compounds of emerging concern in the
Slovene aqueous environment and environmental risk assessment. Environmental Pollution, 242,
143-154.

Chen, D., Kannan, K., Tan, H., Zheng, Z., Feng, Y. L., Wu, Y., Widelka, M. (2016). Bisphenol analogues
other than BPA: environmental occurrence, human exposure, and toxicity: a review. Environmental
science & technology, 50(11), 5438-5453.

Chen Y., Fang J., Ren L., Fan R., Zhang J., Liu G., Zhou L., Chen D., Yu Y., Lu S., (2018) Urinary
bisphenol analogues and triclosan in children from south China and implications for human exposure,
Environmental Pollution 238:299-305.

- Chang, H., Wan, Y., Wu, S., Fan, Z., Hu, J. (2011). Occurrence of androgens and progestogens in
  wastewater treatment plants and receiving river waters: Comparison to estrogens. water research,
  45(2), 732-740.
- 411 Colin, A., Bach, C., Rosin, C., Munoz, J. F., Dauchy, X. (2014). Is drinking water a major route of human
- 412 exposure to alkylphenol and bisphenol contaminants in France? Archives of Environmental 413 Contamination and Toxicology, 66(1), 86-99.
- 414 Corrales, J., Kristofco, L. A., Steele, W. B., Yates, B. S., Breed, C. S., Williams, E. S., Brooks, B. W.
- 415 (2015). Global assessment of bisphenol A in the environment: review and analysis of its occurrence
- 416 and bioaccumulation. Dose-Response, 13(3), 1-29.
- 417 Danzl, E., Sei, K., Soda, S., Ike, M., Fujita, M. (2009). Biodegradation of bisphenol A, bisphenol F and
- 418 bisphenol S in seawater. International journal of environmental research and public health, 6(4),419 1472-1484
- 420 Eladak S., Grisin T., Moison D., Guerquin M-J., N'Tumba-Byn T., Pozzi-Gaudin S., Benachi A., Livera G.,
- 421 Rouiller-Fabre V., Habert R., (2015). A new chapter in the bisphenol A story: bisphenol S and 422 bisphenol F are not safe alternatives to this compound. Fertility and sterility, 103(1), 11-21.
- 423 Halling-Sørensen, B., Nielsen, S. N., Lanzky, P. F., Ingerslev, F., Lützhøft, H. H., Jørgensen, S. (1998).
- 424 Occurrence, fate and effects of pharmaceutical substances in the environment-A review.425 Chemosphere, 36(2), 357-393.
- Hanioka, N., Jinno, H., Tanaka-Kagawa, T., Nishimura, T., & Ando, M. (2000). Interaction of bisphenol
  A with rat hepatic cytochrome P450 enzymes. *Chemosphere*, 41(7), 973-978.
- 428 https://www.drugbank.ca/drugs/DB00472 accessed: November 2018a
- 429 https://www.drugbank.ca/drugs/DB00446 accessed: November 2018b
- 430 https://www.drugbank.ca/drugs/DB00586 accessed: November 2018c
- 431 Integrated Risk Information System (IRIS) U.S. Environmental Protection Agency, Chemical
  432 Assessment Summary, accessed on 27.10.2018 on
  433 https://cfpub.epa.gov/ncea/iris/iris\_documents/documents/subst/0356\_summary.pdf
- Kim, S. D., Cho, J., Kim, I. S., Vanderford, B. J., Snyder, S. A. (2007). Occurrence and removal of
  pharmaceuticals and endocrine disruptors in South Korean surface, drinking, and waste waters.
  Water Research, 41(5), 1013-1021.
- Küster, A., Adler, N. (2014). Pharmaceuticals in the environment: scientific evidence of risks and its
  regulation. Phil. Trans. R. Soc. B, 369(1656), 20130587.
- Li, W. C. (2014). Occurrence, sources, and fate of pharmaceuticals in aquatic environment and soil.
  Environmental Pollution, 187, 193-201.

- Liao, C., Kannan, K. (2013). Concentrations and profiles of bisphenol A and other bisphenol analogues in foodstuffs from the United States and their implications for human exposure. Journal of Agricultural and Food Chemistry, 61(19), 4655-4662.
- Lolić, A., Paíga, P., Santos, L. H., Ramos, S., Correia, M., Delerue-Matos, C. (2015). Assessment of nonsteroidal anti-inflammatory and analgesic pharmaceuticals in seawaters of North of Portugal: occurrence and environmental risk. Science of the Total Environment, 508, 240-250.
- 447 Martínez, C., Ramírez, N., Gómez, V., Pocurull, E., Borrull, F. (2013). Simultaneous determination of 448 76 micropollutants in water samples by headspace solid phase microextraction and gas 449 chromatography-mass spectrometry. Talanta, 116, 937-945.
- Meesters, R. J., Schröder, H. F. (2002). Simultaneous determination of 4-nonylphenol and bisphenol A
  in sewage sludge. Analytical Chemistry, 74(14), 3566-3574.
- 452 Olatunde, O. J., Olatunbosun, O., Olatunde, A., Anyakora, C., Christopher, K. (2014). Occurrence of
  453 selected veterinary pharmaceuticals in water from a fish pond settlement in Ogun State, Nigeria.
  454 International Journal of Environmental Monitoring and Analysis, 2(4), 226.
- Osorio, V., Larrañaga, A., Aceña, J., Pérez, S., Barceló, D. (2016). Concentration and risk of
  pharmaceuticals in freshwater systems are related to the population density and the livestock units
  in Iberian Rivers. Science of the Total Environment, 540, 267-277.
- 458 Owczarek, K., Kudłak, B., Simeonov, V., Mazerska, Z., Namieśnik, J., (2018a). Binary Mixtures of
  459 Selected Bisphenols in the Environment: Their Toxicity in Relationship to Individual Constituents.
  460 Molecules, 23, 3226.
- 461 Owczarek, K., Kubica, P., Kudłak, B., Rutkowska, A., Rachoń, D., Namieśnik, J., Wasik, A. (2018b).
  462 Determination of trace levels of eleven bisphenol A analogues in human blood serum by high
  463 performance liquid chromatography-tandem mass spectrometry. Science of the Total Environment,
  464 628-629, 1362-1368.
- Rocha, M. J., Cruzeiro, C., Rocha, E. (2013). Quantification of 17 endocrine disruptor compounds and
  their spatial and seasonal distribution in the Iberian Ave River and its coastline. Toxicological &
  Environmental Chemistry, 95(3), 386-399.
- 468 Rochester, J. (2013). Bisphenol A and human health: a review on the literature, Reprod. Toxicol, 42,
  469 132-155.
- 470 Rochester, J. R., & Bolden, A. L. (2015). Bisphenol S and F: a systematic review and comparison of the
  471 hormonal activity of bisphenol A substitutes. *Environmental health perspectives*, 123(7), 643-650.
- Sánchez-Avila, J., Fernandez-Sanjuan, M., Vicente, J., Lacorte, S. (2011). Development of a multiresidue method for the determination of organic micropollutants in water, sediment and mussels
  using gas chromatography-tandem mass spectrometry. Journal of Chromatography A, 1218(38),
  6799-6811.

- 476 Sánchez-Avila J., Tauler R., Lacorte, S. (2012). Organic micropollutants in coastal waters from NW
  477 Mediterranean Sea: Sources distribution and potential risk. Environment Internatioanl 46:50-62.
- 478 Santos, L. H., Araújo, A. N., Fachini, A., Pena, A., Delerue-Matos, C., Montenegro, M. C. B. S. M.
- 479 (2010). Ecotoxicological aspects related to the presence of pharmaceuticals in the aquatic480 environment. Journal of hazardous materials, 175(1-3), 45-95.
- 481 Staniszewska, M., Falkowska, L., Grabowski, P., Kwaśniak, J., Mudrak-Cegiołka, S., Reindl, A. R.,
- 482 Sokołowski A., Szumiło E., Zgrundo, A. (2014). Bisphenol A, 4-tert-octylphenol, and 4-nonylphenol in
- the Gulf of Gdańsk (Southern Baltic). Archives of Environmental Contamination and Toxicology, 67(3),
  335-347.
- 485 Stossi, F., Bolt, M. J., Ashcroft, F. J., Lamerdin, J. E., Melnick, J. S., Powell, R. T., Dandekar R.D.,
- 486 Mancini M.D., Walker C.L., Jo Westwick J.K., Mancini, M. A. (2014). Defining estrogenic mechanisms
- 487 of bisphenol A analogs through high throughput microscopy-based contextual assays. *Chemistry* &
  488 *Biology*, 21(6), 743-753.
- 489 Szczepańska, N., Namieśnik, J., Kudłak, B. (2016). Assessment of toxic and endocrine potential of
  490 substances migrating from selected toys and baby products. Environmental Science and Pollution
  491 Research, 23(24), 24890-24900.
- Tanaka Y., Nakamuraa K., Odaa S., Watanabe H., Tatarazako N., (2018). Estimation of populationlevel effect of the endocrine disruptor pyriproxyfen in *Daphnia magna* by using changes in sex ratio
  and reproductive output, Ecotoxicology and Environmental Safety 156, 463–475.
- Wan Y., Huo W., Xu S., Zheng T., Zhang B., Li Y., Zhou A., Zhang Y., Hu J., Zhu Y., Chen Z., Lu S., Wu C.,
  Jiang M., Jiang Y., Liu H., Yang X., Wei Xia (2018). Relationship between maternal exposure to
  bisphenol S and pregnancy duration, Environmental Pollution 238:717-724.
- Weinberger II, J., Klaper, R. (2014). Environmental concentrations of the selective serotonin reuptake
  inhibitor fluoxetine impact specific behaviors involved in reproduction, feeding and predator
  avoidance in the fish Pimephales promelas (fathead minnow). Aquatic toxicology, 151, 77-83.
- Wieczerzak, M., Kudłak, B., Namieśnik, J. (2015). Environmentally oriented models and methods for
  the evaluation of drug× drug interaction effects. Critical Review in Analytical Chemistry, 45, 131-155.)
  Wieczerzak, M., Kudłak, B., Yotova, G., Nedyalkova, M., Tsakovski, S., Simeonov, V., Namieśnik, J.
  (2016). Modeling of pharmaceuticals mixtures toxicity with deviation ratio and best-fit functions
  models. Science of Total Environment, 571, 259-268.
- Wieczerzak, M., Kudłak, B., Yotova, G., Tsakovski, S., Simeonov, V., & Namieśnik, J. (2018a). Impact of
  inorganic ions and pH variations on toxicity and endocrine potential of selected environmentally
  relevant pharmaceuticals. Environmental Pollution, 237, 549-558.
- Xue, J., Liu, W., Kannan, K. (2017). Bisphenols, benzophenones, and bisphenol A diglycidyl ethers in
  textiles and infant clothing. Environmental Science & Technology, 51(9), 5279-5286.

511Yamazaki, E., Yamashita, N., Taniyasu, S., Lam, J., Lam, P. K., Moon, H. B., Jeong Y., Kannan P.,512Achyuthan H., Munuswamy N., Kannan, K. (2015). Bisphenol A and other bisphenol analogues513including BPS and BPF in surface water samples from Japan, China, Korea and India. Ecotoxicology514andEnvironmentalSafety,122,565-572.

Sciected and ased ad	ing rescuren		
	C1	C2	C3
		μM	
BPA	0.548	1.368	2.190
BPS	0.274	0.684	1.095
BPF	0.274	0.684	1.095
Diclofenac s.	5.750	14.375	23.000
Chloramphenicol	8.704	21.759	34.816
Oxytetracycline h.	5.750	14.375	23.000
Fluoxetine h.	8.299	20.748	33.197
Estrone	8.531	21.329	34.127
Ketoprofen	6.865	17.162	27.460
Progesterone	9.516	23.791	38.065
Gemfibrozil	3.736	9.339	14.943
Androstenedione	6.530	16.326	26.121

Table 1. Concentration values of analytes (of bisphenol analogues and pharmaceuticals) selected and used during research

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				YE	S+					YE	S-					YA	\S+					YA	S-		
				М	DR					M	DR					М	DR					M	DR		
		BPA	A C1	BPA	4 C2	BPA	A C3	BPA	C1	BPA	A C2	BPA	\C3	BPA	A C1	BPA	A C2	BPA	A C3	BPA	A C1	BPA	C2	BPA	A C3
		CA	IA																						
	C1	0.76	1.49	0.44	1.12	0.33	0.97	0.92	1.64	0.88	1.63	0.83	1.51	1.09	2.22	0.72	1.51	0.68	1.36	1.07	1.93	0.91	1.60	0.95	1.66
Diclofenac	C2	0.78	1.51	0.50	1.30	0.32	1.02	0.90	1.52	0.88	1.57	0.92	1.59	1.05	1.85	0.96	1.75	0.87	1.57	0.87	1.52	0.90	1.53	0.90	1.50
	C3	0.92	1.81	0.53	1.45	0.36	1.23	0.92	1.56	0.93	1.65	0.98	1.70	1.06	1.82	0.91	1.63	0.89	1.58	0.91	1.57	0.92	1.55	0.94	1.53
Chloramphe-	C1	0.57	1.42	0.37	1.28	0.24	1.00	1.02	1.82	1.03	1.92	0.93	1.70	1.15	2.38	0.84	1.78	0.83	1.69	0.99	1.80	0.93	1.65	0.85	1.47
nicol	C2	1.03	1.98	0.55	1.43	0.30	0.97	0.98	1.72	0.91	1.68	0.92	1.64	1.19	2.27	1.01	1.99	0.82	1.57	0.92	1.66	0.99	1.74	0.94	1.59
Псог	C3	0.73	1.68	0.38	1.26	0.23	0.97	1.12	1.90	1.13	2.00	0.99	1.69	0.98	2.03	0.82	1.76	0.77	1.59	1.18	1.88	1.18	1.86	1.07	1.65
Ovytetracy-	C1	0.56	1.06	0.39	0.93	0.27	0.74	1.08	1.79	0.98	1.72	0.99	1.75	1.51	3.38	1.03	2.33	1.00	2.14	0.99	1.68	0.96	1.62	0.89	1.51
cline h	C2	0.78	1.40	0.37	0.87	0.22	0.62	0.78	1.31	0.91	1.61	0.95	1.63	1.31	2.63	1.03	2.14	1.10	2.21	0.90	1.50	0.95	1.56	0.86	1.39
	C3	0.63	1.17	0.38	0.95	0.24	0.73	1.10	1.73	1.11	1.83	1.10	1.78	1.15	2.30	0.84	1.75	0.97	1.95	0.97	1.59	0.94	1.51	0.91	1.43
	C1	0.66	1.35	0.43	1.18	0.29	0.95	1.14	2.03	1.09	2.01	1.11	2.01	0.74	1.86	0.66	1.67	0.62	1.47	0.86	1.84	0.76	1.54	0.68	1.31
Fluoxetine h.	C2	0.65	1.48	0.41	1.35	0.27	1.12	1.34	2.36	1.22	2.24	1.26	2.23	1.08	2.56	0.99	2.45	0.79	1.85	0.85	2.30	0.75	1.95	0.76	1.81
-	C3	0.77	1.74	0.42	1.35	0.36	1.48	1.20	2.11	1.33	2.45	1.27	2.24	1.13	2.41	0.86	1.91	1.17	2.51	1.19	2.19	1.35	2.42	2.00	3.43
	C1	0.94	1.45	0.79	1.33	0.81	1.46	0.99	1.69	0.88	1.54	0.86	1.59	0.62	0.97	0.45	0.97	0.35	0.96	0.86	1.70	0.81	1.52	0.80	1.47
Estrone	C2	0.88	1.30	0.88	1.38	0.77	1.28	1.08	1.80	0.98	1.62	0.90	1.57	0.74	0.96	0.60	0.96	0.50	0.94	0.91	1.73	0.89	1.59	0.84	1.46
	C3	0.96	1.38	0.96	1.45	0.87	1.40	1.07	1.75	1.02	1.65	0.94	1.59	0.81	1.00	0.67	0.95	0.60	0.96	0.93	1.77	0.80	1.44	0.88	1.50
	C1	0.78	1.52	0.78	1.57	0.67	1.40	0.98	1.95	0.91	1.76	0.89	1.79	0.66	1.50	0.68	1.49	0.68	1.43	0.72	1.45	0.87	1.66	0.62	1.16
Ketoprofen	C2	0.80	1.51	0.74	1.46	0.68	1.43	0.95	1.90	0.89	1.73	0.93	1.90	0.70	1.64	0.70	1.60	0.75	1.66	0.75	1.67	0.76	1.57	0.89	1.75
	C3	0.67	1.22	0.70	1.34	0.69	1.40	1.00	1.87	0.95	1.74	0.97	1.86	0.73	1.63	0.72	1.60	0.62	1.32	0.83	1.77	0.85	1.68	0.81	1.53
	C1	1.06	3.82	0.80	2.92	0.64	2.50	1.09	2.11	1.00	1.89	0.94	1.85	0.61	0.95	0.43	0.93	0.34	0.94	0.82	1.48	0.78	1.36	0.71	1.24
Progesterone	C2	1.07	4.02	0.86	3.39	0.71	3.06	1.04	2.03	0.97	1.85	0.95	1.91	0.66	0.85	0.55	0.89	0.44	0.85	0.80	1.27	0.85	1.31	0.73	1.12
	C3	1.05	3.81	0.82	3.16	0.74	3.11	1.03	1.99	0.95	1.78	1.05	2.08	0.76	0.93	0.59	0.83	0.53	0.87	0.81	1.27	0.77	1.17	0.76	1.13
	C1	1.19	2.41	0.84	1.74	0.81	1.70	1.10	2.24	0.91	1.81	0.84	1.70	1.30	2.93	0.90	1.93	0.87	1.80	0.89	1.73	0.91	1.71	0.95	1.78
Gemfibrozil	C2	0.99	2.14	0.85	1.89	0.79	1.87	0.98	2.11	0.88	1.81	0.85	1.83	1.24	2.35	1.23	2.32	1.16	2.19	0.87	1.70	0.95	1.77	1.00	1.80
	C3	0.94	2.22	0.81	1.98	0.72	1.89	0.96	2.12	0.81	1.71	0.72	1.58	0.94	2.11	0.85	1.87	0.87	1.84	0.88	1.74	0.95	1.77	1.00	1.80
Androstenedi	C1	1.10	1.99	0.85	1.62	0.77	1.53	1.12	2.08	0.97	1.81	0.97	1.87	0.39	0.77	0.29	0.94	0.19	0.85	0.87	1.36	0.77	1.26	0.65	1.16
one	C2	0.97	1.56	1.00	1.70	0.84	1.52	0.99	1.70	0.97	1.66	0.90	1.62	0.63	0.97	0.42	0.96	0.29	0.90	0.86	1.20	0.81	1.18	0.70	1.07
0.10	C3	1.08	1.87	0.89	1.61	0.99	1.90	0.92	1.72	0.87	1.58	0.96	1.84	0.66	0.88	0.55	0.94	0.43	0.92	0.95	1.39	0.92	1.33	0.80	1.17

Table 2. MDR values variations depending on BPA concentration change for solutions of selected pharmaceuticals studied (red - synergism, blue - antagonism, green - overestimation, yellow - underestimation, for values of particular concentrations C1, C2 and C3 of all analytes please refer to Table 1.)

				YE	S+					YE	S-					YA	S+					YA	S-		
				M	DR					М	DR					M	DR					M	DR		
		BPA	\C1	BPA	A C2	BPA	A C3	BPA	C1	BPA	A C2	BPA	A C3	BPA	C1	BPA	C2	BPA	C3	BPA	C1	BPA	C2	BPA	A C3
		CA	IA																						
	C1	0.76	1.49	0.44	1.12	0.33	0.97	0.92	1.64	0.88	1.63	0.83	1.51	1.09	2.22	0.72	1.51	0.68	1.36	1.07	1.93	0.91	1.60	0.95	1.66
Diclofenac	C2	0.78	1.51	0.50	1.30	0.32	1.02	0.90	1.52	0.88	1.57	0.92	1.59	1.05	1.85	0.96	1.75	0.87	1.57	0.87	1.52	0.90	1.53	0.90	1.50
	C3	0.92	1.81	0.53	1.45	0.36	1.23	0.92	1.56	0.93	1.65	0.98	1.70	1.06	1.82	0.91	1.63	0.89	1.58	0.91	1.57	0.92	1.55	0.94	1.53
Chloramphe-	C1	0.57	1.42	0.37	1.28	0.24	1.00	1.02	1.82	1.03	1.92	0.93	1.70	1.15	2.38	0.84	1.78	0.83	1.69	0.99	1.80	0.93	1.65	0.85	1.47
nicol	C2	1.03	1.98	0.55	1.43	0.30	0.97	0.98	1.72	0.91	1.68	0.92	1.64	1.19	2.27	1.01	1.99	0.82	1.57	0.92	1.66	0.99	1.74	0.94	1.59
псог	C3	0.73	1.68	0.38	1.26	0.23	0.97	1.12	1.90	1.13	2.00	0.99	1.69	0.98	2.03	0.82	1.76	0.77	1.59	1.18	1.88	1.18	1.86	1.07	1.65
Ovvtetracy-	C1	0.56	1.06	0.39	0.93	0.27	0.74	1.08	1.79	0.98	1.72	0.99	1.75	1.51	3.38	1.03	2.33	1.00	2.14	0.99	1.68	0.96	1.62	0.89	1.51
cline h	C2	0.78	1.40	0.37	0.87	0.22	0.62	0.78	1.31	0.91	1.61	0.95	1.63	1.31	2.63	1.03	2.14	1.10	2.21	0.90	1.50	0.95	1.56	0.86	1.39
cinic n.	C3	0.63	1.17	0.38	0.95	0.24	0.73	1.10	1.73	1.11	1.83	1.10	1.78	1.15	2.30	0.84	1.75	0.97	1.95	0.97	1.59	0.94	1.51	0.91	1.43
	C1	0.66	1.35	0.43	1.18	0.29	0.95	1.14	2.03	1.09	2.01	1.11	2.01	0.74	1.86	0.66	1.67	0.62	1.47	0.86	1.84	0.76	1.54	0.68	1.31
Fluoxetine h.	C2	0.65	1.48	0.41	1.35	0.27	1.12	1.34	2.36	1.22	2.24	1.26	2.23	1.08	2.56	0.99	2.45	0.79	1.85	0.85	2.30	0.75	1.95	0.76	1.81
	C3	0.77	1.74	0.42	1.35	0.36	1.48	1.20	2.11	1.33	2.45	1.27	2.24	1.13	2.41	0.86	1.91	1.17	2.51	1.19	2.19	1.35	2.42	2.00	3.43
	C1	0.94	1.45	0.79	1.33	0.81	1.46	0.99	1.69	0.88	1.54	0.86	1.59	0.62	0.97	0.45	0.97	0.35	0.96	0.86	1.70	0.81	1.52	0.80	1.47
Estrone	C2	0.88	1.30	0.88	1.38	0.77	1.28	1.08	1.80	0.98	1.62	0.90	1.57	0.74	0.96	0.60	0.96	0.50	0.94	0.91	1.73	0.89	1.59	0.84	1.46
	C3	0.96	1.38	0.96	1.45	0.87	1.40	1.07	1.75	1.02	1.65	0.94	1.59	0.81	1.00	0.67	0.95	0.60	0.96	0.93	1.77	0.80	1.44	0.88	1.50
	C1	0.78	1.52	0.78	1.57	0.67	1.40	0.98	1.95	0.91	1.76	0.89	1.79	0.66	1.50	0.68	1.49	0.68	1.43	0.72	1.45	0.87	1.66	0.62	1.16
Ketoprofen	C2	0.80	1.51	0.74	1.46	0.68	1.43	0.95	1.90	0.89	1.73	0.93	1.90	0.70	1.64	0.70	1.60	0.75	1.66	0.75	1.67	0.76	1.57	0.89	1.75
	C3	0.67	1.22	0.70	1.34	0.69	1.40	1.00	1.87	0.95	1.74	0.97	1.86	0.73	1.63	0.72	1.60	0.62	1.32	0.83	1.77	0.85	1.68	0.81	1.53
	C1	1.06	3.82	0.80	2.92	0.64	2.50	1.09	2.11	1.00	1.89	0.94	1.85	0.61	0.95	0.43	0.93	0.34	0.94	0.82	1.48	0.78	1.36	0.71	1.24
Progesterone	C2	1.07	4.02	0.86	3.39	0.71	3.06	1.04	2.03	0.97	1.85	0.95	1.91	0.66	0.85	0.55	0.89	0.44	0.85	0.80	1.27	0.85	1.31	0.73	1.12
	C3	1.05	3.81	0.82	3.16	0.74	3.11	1.03	1.99	0.95	1.78	1.05	2.08	0.76	0.93	0.59	0.83	0.53	0.87	0.81	1.27	0.77	1.17	0.76	1.13
	C1	1.19	2.41	0.84	1.74	0.81	1.70	1.10	2.24	0.91	1.81	0.84	1.70	1.30	2.93	0.90	1.93	0.87	1.80	0.89	1.73	0.91	1.71	0.95	1.78
Gemfibrozil	C2	0.99	2.14	0.85	1.89	0.79	1.87	0.98	2.11	0.88	1.81	0.85	1.83	1.24	2.35	1.23	2.32	1.16	2.19	0.87	1.70	0.95	1.77	1.00	1.80
	C3	0.94	2.22	0.81	1.98	0.72	1.89	0.96	2.12	0.81	1.71	0.72	1.58	0.94	2.11	0.85	1.87	0.87	1.84	0.88	1.74	0.95	1.77	1.00	1.80
Androstanadi	C1	1.10	1.99	0.85	1.62	0.77	1.53	1.12	2.08	0.97	1.81	0.97	1.87	0.39	0.77	0.29	0.94	0.19	0.85	0.87	1.36	0.77	1.26	0.65	1.16
one	C2	0.97	1.56	1.00	1.70	0.84	1.52	0.99	1.70	0.97	1.66	0.90	1.62	0.63	0.97	0.42	0.96	0.29	0.90	0.86	1.20	0.81	1.18	0.70	1.07
one	C3	1.08	1.87	0.89	1.61	0.99	1.90	0.92	1.72	0.87	1.58	0.96	1.84	0.66	0.88	0.55	0.94	0.43	0.92	0.95	1.39	0.92	1.33	0.80	1.17

Table 2. MDR values variations depending on BPA concentration change for solutions of selected pharmaceuticals studied (legend for printed B&W version: MDR > 2.0 exhibits antagonism, MDR < 0.5 shows synergism, MDR values within 0.50-0.71 and 1.40-2.00 values mean, respectively, under- and overestimation of presented models, for values of particular concentrations C1, C2 and C3 of all analytes please refer to Table 1.)

Bisphenols (A, S, and F) affect the basic hormonal activity determined for pharmaceuticals – study of Saccharomyces cerevisiae

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## Electronic supplementary material

				YF	S+		inaryces			YF	·ς_					YΔ	S+					YΔ	S-		
				MI						M	DR					M	DR					M	DR		
		BPS	C1	BPS	5.02	BPS	5.03	BPS	C1	BPS	5.02	BPS	C3	BPS	C1	BPS	5.02	BPS	5.03	BPS	C1	BPS	C2	BPS	C3
		CA	IA	CA	IA	CA	IA	CA	IA	CA	IA	CA	IA	CA	IA	CA	IA	CA	IA	CA	IA	CA	IA	CA	IA
	C1	0.69	1.46	0.49	1.40	0.22	0.95	1.17	2.43	0.91	2.14	0.67	1.48	1.21	2.22	0.63	1.28	0.63	1.27	0.90	1.45	0.91	1.48	0.78	1.27
Diclofenac	C2	0.77	1.52	0.49	1.35	0.28	1.19	0.97	2.06	0.89	2.20	0.56	1.32	0.74	1.29	0.70	1.36	0.56	1.08	0.97	1.51	0.91	1.45	0.85	1.36
	C3	0.73	1.49	0.51	1.46	0.29	1.31	1.29	2.57	0.97	2.25	0.67	1.47	0.78	1.39	0.73	1.43	0.69	1.34	0.90	1.41	0.88	1.41	0.87	1.39
	C1	0.80	1.69	0.46	1.35	0.23	1.05	1.06	2.37	0.94	2.42	0.52	1.26	1.05	1.93	0.82	1.68	0.69	1.37	0.97	1.55	0.96	1.56	0.81	1.31
Chloramphe	C2	0.93	1.83	0.60	1.64	0.25	1.10	1.06	2.13	0.85	1.98	0.63	1.40	1.02	1.77	0.94	1.82	0.76	1.44	1.05	1.63	0.97	1.55	0.90	1.43
-nicoi	C3	0.84	1.70	0.49	1.40	0.24	1.09	1.17	2.29	0.94	2.12	0.66	1.43	1.04	1.87	0.96	1.92	0.67	1.31	0.93	1.46	0.93	1.49	0.90	1.44
<u> </u>	C1	1.04	2.02	0.49	1.27	0.21	0.82	1.08	2.09	0.74	1.61	0.66	1.38	0.88	1.68	0.67	1.42	0.68	1.39	0.93	1.50	0.87	1.42	0.92	1.50
Oxytetracy-	C2	0.94	1.86	0.46	1.26	0.29	1.25	1.08	1.93	0.74	1.49	0.75	1.46	0.90	1.59	0.79	1.55	0.76	1.47	0.96	1.51	0.99	1.58	0.93	1.48
cine n.	C3	0.84	1.55	0.38	0.97	0.26	1.01	0.89	1.64	0.62	1.30	0.61	1.23	0.99	1.66	0.77	1.44	0.79	1.45	0.97	1.51	0.97	1.54	0.88	1.40
El time	C1	0.80	1.59	0.53	1.45	0.27	1.16	0.99	2.05	0.63	1.49	0.63	1.41	0.84	1.63	0.75	1.61	0.69	1.47	0.88	1.60	0.80	1.47	0.82	1.49
Fluoxetine	C2	0.82	1.67	0.47	1.34	0.31	1.41	0.82	1.75	0.60	1.49	0.59	1.40	0.74	1.71	0.74	1.94	0.77	1.97	0.84	2.04	0.74	1.84	0.77	1.89
11.	C3	0.79	1.67	0.47	1.43	0.26	1.28	0.83	1.66	0.56	1.29	0.73	1.62	0.96	2.16	1.01	2.60	0.98	2.45	1.05	2.57	1.19	3.02	1.33	3.35
	C1	0.82	1.20	0.76	1.12	0.82	1.26	1.02	1.42	0.99	1.53	0.98	1.54	0.69	0.93	0.42	0.68	0.37	0.75	0.90	1.58	0.74	1.33	0.72	1.21
Estrone	C2	1.00	1.22	0.83	1.08	0.94	1.30	1.09	1.44	1.00	1.47	1.13	1.66	0.76	0.92	0.52	0.68	0.49	0.74	0.90	1.64	0.65	1.25	0.79	1.38
	C3	1.02	1.19	0.89	1.12	0.94	1.29	0.87	1.68	0.64	1.52	0.79	1.79	0.85	0.98	0.54	0.67	0.56	0.76	0.84	1.27	0.72	1.12	0.85	1.26
	C1	0.67	1.12	0.30	0.87	0.25	0.94	0.81	1.54	0.57	1.31	0.67	1.46	0.95	1.84	0.95	1.84	0.69	1.36	0.82	1.34	0.81	1.35	0.69	1.08
Ketoprofen	C2	0.91	1.43	0.31	0.83	0.28	0.99	0.78	1.60	0.59	1.49	0.76	1.86	0.89	1.81	0.89	1.81	0.68	1.40	0.83	1.44	0.59	1.06	0.70	1.18
	C3	0.83	1.14	0.39	0.82	0.48	1.29	0.86	1.58	0.66	1.48	0.74	1.61	0.98	1.93	0.98	1.93	0.69	1.38	0.74	1.31	0.63	1.17	0.78	1.32
Progesteron	C1	0.86	2.78	0.30	2.33	0.18	1.96	1.00	2.48	0.64	1.99	0.65	1.90	0.56	0.81	0.41	0.73	0.30	0.74	0.77	1.50	0.69	1.40	0.67	1.24
A	C2	0.65	2.04	0.32	2.40	0.17	1.87	0.93	2.15	0.76	2.22	0.72	2.02	0.72	0.89	0.56	0.79	0.52	0.92	0.78	1.41	0.65	1.22	0.72	1.24
C	C3	0.67	1.96	0.31	2.16	0.22	2.18	0.91	2.11	0.72	2.14	0.70	1.97	0.73	0.84	0.68	0.86	0.62	0.92	0.81	1.45	0.74	1.39	0.83	1.44
	C1	1.24	2.17	0.39	1.14	0.17	0.65	1.30	2.47	0.77	1.72	0.90	1.90	1.69	3.27	0.56	1.21	0.66	1.31	0.85	1.37	0.82	1.34	0.82	1.27
Gemfibrozil	C2	1.79	2.83	0.42	1.11	0.23	0.80	1.10	2.26	0.69	1.73	0.81	1.92	0.89	1.82	0.55	1.26	0.58	1.18	0.74	1.29	0.72	1.29	0.77	1.28
	C3	1.24	1.98	0.26	0.71	0.34	1.23	1.48	2.74	0.87	1.95	1.10	2.36	0.87	1.72	0.49	1.11	0.69	1.38	0.83	1.45	0.69	1.26	0.79	1.34
Androstene-	C1	1.60	2.23	0.76	1.38	0.32	0.68	0.94	1.77	0.63	1.42	0.71	1.52	0.49	0.94	0.34	0.88	0.22	0.87	0.94	1.68	0.89	1.63	0.88	1.48
dione	C2	3.22	4.00	1.15	1.58	0.74	1.10	0.83	1.39	0.72	1.43	0.70	1.34	0.67	0.95	0.53	0.93	0.39	0.97	0.84	1.43	0.86	1.53	0.81	1.34
	C3	2.91	3.48	0.58	0.84	0.68	1.12	1.18	2.15	0.64	1.42	0.89	1.90	0.70	0.88	0.62	0.91	0.46	0.89	0.96	1.60	0.92	1.60	0.88	1.43

Table A. MDR values variations depending on BPS concentration change for solutions of selected pharmaceuticals studied (red - synergism, blue - antagonism, green - overestimation, yellow – underestimation, for values of particular concentrations C1, C2 and C3 of all analytes please refer to Table 1.)

## Electronic supplementary material

Table B. MDR values variations depending on BPF concentration change for solutions of selected pharmaceuticals studied (red - synergism, blue - antagonism, green - overestimation, yellow - underestimation, for values of particular concentrations C1, C2 and C3 of all analytes please refer to Table 1.)

				YE	S+					YE	ES-					YA	\S+					YA	\S-		
				Μ	DR					М	DR					М	DR					M	DR		
		BPF	C1	BP	F C2	BPF	- C3	BPF	C1	BPF	C2	BPF	C3	BPF	C1	BP	FC2	BPF	F C3	BPF	C1	BPF	C2	BPF	C3
		CA	IA																						
	C1	0.77	1.98	0.33	1.39	0.27	1.15	1.02	2.24	0.69	1.57	0.66	1.52	1.26	2.57	0.73	1.45	0.65	1.32	1.31	2.55	1.14	2.12	1.05	1.94
Diclofenac	C2	0.65	1.77	0.29	1.34	0.26	1.27	0.91	1.96	0.74	1.65	0.73	1.71	0.84	1.69	0.71	1.39	0.56	1.13	0.99	1.96	0.97	1.81	0.95	1.75
	C3	0.70	1.91	0.31	1.49	0.28	1.38	0.95	2.07	0.78	1.76	0.74	1.75	0.97	1.63	0.81	1.35	0.76	1.30	1.01	1.88	0.91	1.59	0.91	1.59
Chloramphe-	C1	0.77	1.94	0.38	1.57	0.29	1.25	1.31	2.53	0.88	1.76	0.82	1.70	1.17	2.48	0.68	1.40	0.61	1.27	0.86	1.61	0.75	1.32	0.74	1.27
nicol	C2	0.76	1.91	0.28	1.18	0.29	1.31	1.13	2.18	0.91	1.82	0.90	1.88	0.74	1.59	0.63	1.31	0.49	1.06	0.47	1.26	0.47	1.15	0.46	1.08
	C3	0.84	2.03	0.42	1.69	0.31	1.34	1.28	2.38	1.04	2.01	0.98	1.99	0.76	1.42	0.64	1.17	0.60	1.13	0.43	1.19	0.42	1.05	0.42	1.04
Oxytetracy-	C1	0.77	1.72	0.30	1.04	0.27	0.96	0.93	1.84	0.73	1.49	0.66	1.39	0.96	1.81	0.85	1.57	0.73	1.40	0.97	1.54	0.94	1.44	0.81	1.25
cline h.	C2	0.72	1.60	0.29	1.05	0.25	0.93	0.92	1.68	0.86	1.63	0.74	1.46	0.92	1.65	0.88	1.55	0.69	1.25	0.98	1.49	0.99	1.44	0.96	1.39
	C3	0.79	1.70	0.32	1.11	0.30	1.09	1.18	2.10	0.76	1.40	0.82	1.57	1.00	1.69	0.93	1.55	0.85	1.46	1.04	1.58	0.99	1.44	0.91	1.32
	C1	0.92	1.93	0.37	1.22	0.31	1.05	1.04	2.00	0.83	1.66	0.77	1.59	0.71	1.50	0.74	1.52	0.69	1.44	0.82	1.53	0.82	1.44	0.79	1.36
Fluoxetine h.	C2	0.97	1.98	0.39	1.27	0.37	1.26	0.91	1.76	0.90	1.79	0.83	1.71	0.82	1.77	0.84	1.75	0.81	1.75	0.85	2.26	0.81	1.97	0.89	2.10
	C3	0.80	1.71	0.35	1.22	0.32	1.14	0.94	1.75	0.87	1.69	0.77	1.56	0.83	1.55	0.89	1.62	0.95	1.78	0.68	1.87	0.79	1.97	0.97	2.36
	C1	0.78	1.15	0.75	1.15	0.68	1.17	1.10	1.50	0.78	1.15	0.80	1.27	0.62	0.89	0.43	0.85	0.36	0.82	0.87	1.55	0.84	1.53	0.80	1.40
Estrone	C2	0.96	1.18	0.94	1.24	0.77	1.08	1.00	1.33	0.82	1.10	0.81	1.19	0.77	0.93	0.59	0.84	0.55	0.86	1.02	1.85	0.92	1.72	0.90	1.61
	C3	1.02	1.19	0.90	1.14	0.77	1.01	1.17	1.49	1.00	1.29	1.02	1.42	0.68	0.79	0.66	0.85	0.60	0.83	0.92	1.77	0.79	1.55	0.78	1.48
	C1	0.60	0.92	0.43	0.91	0.35	0.75	0.95	1.49	0.67	1.07	0.69	1.22	0.74	1.17	0.70	1.16	0.68	1.18	0.80	1.28	0.82	1.34	0.79	1.26
Ketoprofen	C2	0.93	1.46	0.52	1.19	0.49	1.13	0.69	1.33	0.47	0.88	0.58	1.28	0.71	1.36	0.62	1.18	0.56	1.13	0.85	1.41	0.77	1.31	0.74	1.23
	C3	0.66	0.94	0.54	1.07	0.44	0.89	0.89	1.50	0.68	1.11	0.64	1.22	0.63	1.11	0.63	1.11	0.76	1.43	0.79	1.27	0.80	1.31	0.82	1.31
	C1	0.91	2.41	0.42	2.00	0.35	1.66	1.13	2.39	0.74	1.49	0.65	1.55	0.58	0.84	0.40	0.80	0.35	0.82	0.78	1.60	0.63	1.32	0.70	1.39
Progesterone	C2	0.83	2.21	0.39	1.92	0.35	1./3	1.07	2.13	0.82	1.57	0.75	1.72	0.68	0.82	0.55	0.80	0.50	0.81	0.89	1.55	0.87	1.56	0.87	1.51
	C3	0.83	2.12	0.39	1.8/	0.32	1.54	0.92	1.78	0.76	1.41	0.80	1.79	0.74	0.86	0.59	0.78	0.65	0.92	0.89	1.59	0.77	1.41	0.91	1.62
Constitution and	C1	0.90	2.18	0.26	1.14	0.20	0.89	1.13	2.21	0.60	1.14	0.62	1.32	0.89	1.69	0.64	1.23	0.48	0.96	0.75	1.22	0.77	1.28	0.73	1.19
Gemfibrozii	C2	0.69	1.88	0.27	1.38	0.22	1.16	0.90	2.19	0.55	1.26	0.54	1.49	0.71	1.47	0.61	1.26	0.53	1.15	0.86	1.42	0.80	1.36	0.79	1.30
	C3	0.59	1.55	0.27	1.33	0.22	1.08	0.94	2.05	0.00	1.36	0.70	1.76	0.72	1.43	0.62	1.22	0.65	1.37	0.75	1.24	0.76	1.30	0.92	1.53
Androstene-		1.49	2.27	0.68	1.42	0.50	1.04	1.33	2.20	0.80	1.33	0.75	1.39	0.57	1.06	0.34	1.02	0.27	0.96	0.88	1.41	0.87	1.42	0.83	1.34
dione	C2	0.85	1.23	0.44	0.88	0.37	0.75	1.08	1.53	0.85	1.23	0.78	1.24	0.0/	0.91	0.49	0.91	0.40	0.86	0.97	1.50	0.92	1.46	0.85	1.32
	C3	1.56	2.21	0.70	1.36	0.67	1.33	1.06	1.73	0.81	1.30	0.83	1.53	0.75	0.93	0.60	0.92	0.53	0.93	0.97	1.52	0.96	1.54	1.00	1.56

Table C. Percentile values for model deviation ratios (MDR) and numbers of cases for each group of CA and IA experiments of BPA YES+ toxicity studies.

	80%	90%	95%	99%	Synergism	Under- estimation	Over- estimation	Antagon ism	Total
CA	0.89	0.99	1.06	1.11	21	14	0	0	81
IA	1.89	2.50	3.16	3.86	0	1	34	12	81

Table D. Percentile values for model deviation ratios (MDR) and numbers of cases for each group of CA and IA experiments of BPA YES- toxicity studies.

	80%	90%	95%	99%	Synergism	Under- estimation	Over- estimation	Antago- nism	Total
CA	1.09	1.12	1.22	1.34	0	0	0	0	81
IA	2.00	2.11	2.24	2.38	0	0	63	17	81

Table E. Percentile values for model deviation ratios (MDR) and numbers of cases for each group of CA and IA experiments of BPA YAS+ toxicity studies.

	80%	90%	95%	99%	Synergism	Under- estimation	Over- estimation	Antago- nism	Total
CA	1.03	1.15	1.23	1.35	12	21	1	0	81
IA	2.14	2.35	2.51	3.02	0	0	32	20	81

Table F. Percentile values for model deviation ratios (MDR) and numbers of cases for each group of CA and IA experiments of BPA YAS- toxicity studies.

	80%	90%	95%	99%	Synergism	Under- estimation	Over- estimation	Antago- nism	Total
CA	0.95	1.00	1.18	1.48	0	4	1	0	81
IA	1.77	1.84	1.95	2.62	0	0	57	4	81

Table G. Percentile values for model deviation ratios (MDR) and numbers of cases for each group of CA and IA experiments of BPS toxicity studies.

	80%	90%	95%	99%	Synergism	Under- estimation	Over- estimation	Antago- nism	Total
CA	0.86	1.02	1.24	2.97	40	9	2	2	81
IA	1.86	2.17	2.40	3.59	0	3	22	12	81

Table H. Percentile values for model deviation ratios (MDR) and numbers of cases for each group of CA and IA experiments of BPS YES- toxicity studies.

	80%	90%	95%	99%	Synergism	Under- estimation	Over- estimation	Antago- nism	Total
CA	1.00	1.10	1.17	1.34	0	27	1	0	81
IA	2.14	2.29	2.43	2.61	0	0	46	24	81

Table I. Percentile values for model deviation ratios (MDR) and numbers of cases for each group of CA and IA experiments of BPS YAS+ toxicity studies.

	80%	90%	95%	99%	Synergism	Under- estimation	Over- estimation	Antago- nism	Total
CA	0.88	0.98	1.02	1.30	11	34	1	0	81
IA	1.72	1.93	2.16	2.73	0	3	31	5	81

	80%	90%	95%	99%	Synergism	Under- estimation	Over- estimation	Antago- nism	Total
CA	0.93	0.97	0.99	1.22	0	9	0	0	81
IA	1.55	1.63	1.89	3.08	0	0	43	4	81

Table J. Percentile values for model deviation ratios (MDR) and numbers of cases for each group of CA and IA experiments of BPS YAS- toxicity studies.

Table K. Percentile values for model deviation ratios (MDR) and numbers of cases for each group of CA and IA experiments of BPF YES+ toxicity studies.

	80%	90%	95%	99%	Synergism	Under- estimation	Over- estimation	Antago- nism	Total
CA	0.79	0.91	0.96	1.50	43	12	2	0	81
IA	1.87	1.98	2.18	2.30	0	0	23	8	81

Table L. Percentile values for model deviation ratios (MDR) and numbers of cases for each group of CA and IA experiments of BPF YES- toxicity studies.

		80%	90%	95%	99%	Synergism	Under- estimation	Over- estimation	Antago- nism	Total
	CA	1.00	1.10	1.17	1.31	1	16	0	0	81
ſ	IA	1.88	2.13	2.21	2.41	0	0	44	14	81

Table M. Percentile values for model deviation ratios (MDR) and numbers of cases for each group of CA and IA experiments of BPF YAS+ toxicity studies.

	80%	90%	95%	99%	Synergism	Under- estimation	Over- estimation	Antago- nism	Total
CA	0.82	0.89	0.96	1.19	10	39	0	0	81
IA	1.55	1.69	1.77	2.49	0	0	26	2	81

Table N. Percentile values for model deviation ratios (MDR) and numbers of cases for each group of CA and IA experiments of BPF YAS- toxicity studies.

	80%	90%	95%	99%	Synergism	Under- estimation	Over- estimation	Antago- nism	Total
CA	0.96	0.99	1.02	1.18	6	3	0	0	81
IA	1.62	1.94	2.10	2.40	0	0	45	5	81