Identification of novel halogenated naturally occurring compounds in marine biota by high-resolution mass spectrometry and combined screening approaches

Lidia Belova, Yukiko Fujii, Paulien Cleys, Monika Śmiełowska, Koichi Haraguchi, Adrian Covaci

PII: S0269-7491(21)01515-3

DOI: https://doi.org/10.1016/j.envpol.2021.117933

Reference: ENPO 117933

To appear in: Environmental Pollution

Received Date: 17 June 2021

Revised Date: 4 August 2021

Accepted Date: 5 August 2021

Please cite this article as: Belova, L., Fujii, Y., Cleys, P., Śmiełowska, M., Haraguchi, K., Covaci, A., Identification of novel halogenated naturally occurring compounds in marine biota by high-resolution mass spectrometry and combined screening approaches, *Environmental Pollution* (2021), doi: https://doi.org/10.1016/j.envpol.2021.117933.

This is a PDF file of an article that has undergone enhancements after acceptance, such as the addition of a cover page and metadata, and formatting for readability, but it is not yet the definitive version of record. This version will undergo additional copyediting, typesetting and review before it is published in its final form, but we are providing this version to give early visibility of the article. Please note that, during the production process, errors may be discovered which could affect the content, and all legal disclaimers that apply to the journal pertain.

© 2021 Published by Elsevier Ltd.



Credit author statement

Lidia Belova: Conceptualization, Methodology, Formal analysis, Investigation, Writing – original draft preparation, Visualization

Yukiko Fujii: Conceptualization, Methodology, Formal analysis, Investigation, Writing – original

draft preparation, Visualization, Funding acquisition

Paulien Cleys: Methodology, Formal analysis, Investigation, Writing – original draft preparation
 Monika Śmiełowska: Methodology, Formal analysis, Investigation, Writing – review and editing
 Koichi Haraguchi: Conceptualization, Writing – review and editing

Adrian Covaci: Conceptualization, Methodology, Resources, Writing – review and editing, Supervision, Project Administration, Funding acquisition



Naturally occurring halogenated compounds (NHCs)

Journal Prevent

ournal	Dra nroo	
oumai		

	15
ld	16
edzy.	17
ostwi	18
m mo	19
ded fr	20
vnloa	21
Dov	22
λZ	3
	4
S F	5
MOS	6
$\langle $	

Identification of novel halogenated naturally occurring compounds in marine biota by high-resolution mass spectrometry and combined screening approaches 3

Lidia Belova^{1,#}, Yukiko Fujii^{1,2,#}, Paulien Cleys¹, Monika Śmiełowska³, Koichi Haraguchi², Adrian
 Covaci^{1,*}

6

7

1 - Toxicological Centre, University of Antwerp, Universiteitsplein 1, 2610, Antwerp, Belgium

8 2 - Department of Pharmaceutical Sciences, Daiichi University of Pharmacy, 815–8511, Fukuoka, Japan

9 3 - Department of Analytical Chemistry, Gdańsk University of Technology, 80-233, Gdańsk, Poland

10

11 [#]-shared first authors

12 *-corresponding author

13

14 Abstract:

Marine animals, plants or bacteria are a source of bioactive naturally-occurring halogenated compounds 15 (NHCs) such as bromophenols (BPs), bromoanisoles (BAs) and hydroxylated or methoxylated analogues of polybrominated diphenyl ethers (HO-PBDEs, MeO-PBDEs) and bromobiphenyls (HO-BBs, MeO-BBs). This study applied a comprehensive screening approach using liquid chromatography 18 high-resolution mass spectrometry and combining target, suspect and non-target screening with the aim to identify new hydroxylated NHCs which might be missed by commonly applied gas chromatographic methods. 24 alga samples, 4 sea sponge samples and 7 samples of other invertebrates were screened. Target screening was based on 19 available reference standards of BPs, (di)OH-BDEs and diOH-BBs 3 and yielded seven unequivocally identified compounds. 6OH-BDE47 was the most frequently detected compound with a detection frequency of 31%. Suspect screening yielded two additional compounds 4 5 identified in alga samples as well as 17 and 8 compounds identified in sea sponge samples of 6 Lamellodysidea sp. and Callyspongia sp., respectively. The suspect screening results presented here

27 confirmed the findings of previous studies conducted on sea sponge samples of Lamellodysidea sp. and Callyspongia sp. Additionally, in Lamellodysidea sp. and Callyspongia sp. 13 and 4 newly identified 28 29 NHCs are reported including heptabrominated diOH-BDE, monochlorinated pentabrominated 30 diOH-BDE, hexabrominated OH-MeO-BDE and others. Non-target screening allowed the identification of 31 and 20 polyhalogenated compounds in Lamellodysidea sp. and Callyspongia sp. samples, 31 32 respectively. Based on the obtained fragmentation spectra, polybrominated dihydroxylated diphenoxybenzenes (diOH-PBDPBs), such as hepta-, octa- and nonabrominated diOH-BDPBs, could 33 be identified in both species. To our knowledge, this study is the first report on the environmental 34 presence of OH-PBDPBs. 35

36

Keywords: naturally-occurring halogenated compounds, seafood, sea sponges, HO-BDE,
 polybrominated dihydroxylated diphenoxybenzenes, target screening, suspect and non-target screening

Introduction: 40

Natural halogenated compounds (NHCs) are a diverse class of compounds commonly detectable in the 41 marine environment and produced by marine animals, plants or bacteria (Gribble, 2003). The most 42 43 relevant NHC classes found in biota include bromophenols (BPs), bromoanisoles (BAs) and hydroxylated or methoxylated analogues of brominated biphenyls (HO-BBs and MeO-BBs), of 44 polybrominated diphenyl ethers (HO-PBDEs and MeO-PBDEs), and of brominated dibenzo-p-dioxins 45 46 (HO-BDDs and MeO-BDDs) (Bidleman et al., 2019a; Bidleman et al., 2019b). These brominated 47 compounds may originate both from natural and anthropogenic sources. For example, several BPs are used as brominated flame retardants (BFRs) and precursors for the production of more complex BFRs 48 (Thomsen et al., 2001), but also may occur in red, green and brown algae, as well as sea sponges 49 50 (Haraguchi et al., 2010; Liu et al., 2011). HO-PBDEs are metabolites of anthropogenic PBDEs, but can 51 also have a natural origin e.g., when produced by marine bacteria (Agarwal et al., 2014). MeO-PBDEs are presumed to be of natural origin only (Hamers et al., 2008; Wan et al., 2009). BDDs are produced 52 as by-products in the industrial processing or thermal decomposition of BFRs, but have also been 53 described to occur in red alga (Malmvarn et al., 2008). 54

The resistance to chemical and biological degradation of natural and anthropogenic halogenated compounds can lead to their bioaccumulation and biomagnification in the environment. The bioaccumulation and biomagnification potential of PBDEs and their analogies was studied in different aquatic food chains by determining the concentration of PBDEs, HO-PBDEs and MeO-PBDEs in mammals, fish and other invertebrates which represented different trophic levels (Dahlgren et al., 2016; Kim et al., 2015; Sun et al., 2020; Weijs et al., 2009). Recent studies showed that the bioaccumulative 2 potential is highest for MeO-PBDEs, followed by PBDEs and HO-PBDEs, respectively (Losada et al., 3 2009; Sun et al., 2020). This in turn can lead to biomagnification of NHCs in higher trophic levels of the food chain (Choo et al., 2019). Due to this fact and the structural similarities of NHCs to man-made toxic 4 environmental pollutants (e.g., BFRs), potential toxic effects of NHCs have been a topic of interest during

5

55

the last decade (Dingemans et al., 2008; Linares et al., 2015). For example, described toxic effects of
OH-PBDEs included disruption of thyroid hormone homeostasis, oxidative phosphorylation disruption,
neurotoxic effects and altered estradiol synthesis showing that the investigation of hydroxylated NHCs
is of high relevance to access possible toxic effects on humans (Dingemans et al., 2008; Li et al., 2010;
Su et al., 2014; Van Boxtel et al., 2007).

71

Most studies on characterization and quantification of NHCs are carried out applying gas 72 73 chromatography (GC) coupled to tandem mass spectrometry (MS/MS) or high-resolution mass 74 spectrometry (HRMS). Thereby, the separation of neutral and hydroxylated NHCs is carried out by fractionation within sample preparation followed by derivatization of hydroxylated NHCs to methoxylated 75 76 derivates prior to analysis (Lacorte et al., 2010; Malmvarn et al., 2005). While this approach shows great 77 selectivity for the detection and identification of methoxylated NHCs, the identification of e.g. HO-MeO-PBDEs or HO-MeO-BBs can be hampered by the presence of corresponding dihydroxy analogues and 78 79 vice versa (Kato et al., 2012).

80 To address this issue, several methods using liquid chromatography (LC) coupled HRMS or MS/MS for 81 the analysis of hydroxylated NHCs have been introduced in the past (Kato et al., 2012; Lupton et al., 82 2010; Wang et al., 2011). Thereby, the application of targeted methods allows the tentative identification 83 (and quantification) of known NHCs providing valuable information to access potential human exposure 84 (Kato et al., 2009). However, this approach leaves many unknown potentially bioactive or toxic NHCs undetected. The use of HRMS for suspect or non-target screening (NTS) approaches allows the 85 simultaneous analysis of a high number of compounds. Thereby, the application of novel bioinformatics 86 87 tools, such as HaloSeeker, enables a selective detection of halogenated compounds based on their 8 specific isotopic patterns (Leon et al., 2019). Previous studies applying NTS for the identification of novel 9 NHCs in biota samples have shown the potential of these techniques (Agarwal et al., 2015; Cariou et 0 al., 2020). Nevertheless, a study combining the advantages of both targeted and non-target approaches using LC-HRMS for the detection and identification of HO-NHCs is still lacking.

Therefore, we aim to comprehensively screen a variety of biota samples using a combination of target, 92 suspect and non-target screening approaches with the purpose to simultaneously detect known NHCs 93 94 and potentially discover new.

95

Materials and methods 96

Chemicals 97

HPLC-grade methanol and acetonitrile were obtained from Biosolve Chimie SARL (Dieuze, France). A 98 99 PURELAB Flexsystem was used to obtain ultrapure water (18.2 MΩ cm, Milli-Q, Millipore). Ammonium acetate was purchased from Sigma-Aldrich (eluent additive for LC-MS). Solvents used for sample 100 preparation included ethyl acetate, methanol, isooctane, dichloromethane, diethyl ether, sodium 101 102 sulphate (all from Merck KGaA, Darmstadt, Germany), hexane (Acros Organics, Geel, Belgium), 1M 103 KOH, ethanol (EtOH) and 1M hydrogen chloride (HCl). 4-hydroxy-polychlorinated biphenyl-159 (4-OH-PCB159) was used as internal standard and d18-gamma-hexabromocyclododecane (d18-y-HBCD) as 104 recovery standard, respectively. Both standards were provided by Wellington Laboratories, Canada. 105 106 Standards available for targeted screening included 14 OH-PBDEs, 19 BPs, 2,2'-diOH-BB80 and 2,2'-107 diOH-BDE68. The exact compounds including their full name, molecular formula and commercial source **801 10 10 10 10 10 111** are summarized in Table S1. Stock solutions were available at concentrations of 10 µg/mL or 50 µg/mL in acetonitrile. The same solvent was used to prepare working solutions at a concentration of 5 µg/mL and 0.5 µg/mL, which were stored at -20 °C prior to analysis.

4

5

6

7

0

MOST WIEDZY

Samples

A total of 35 samples were collected from the natural habitat of the species or purchased from the commercial market. For each sample, the sample identification, the sample type, the species, the edibility and the origin are shown in Table S2. Sample 31 was a stored extract which was prepared by the sample preparation method described in Kato et al. (2012). The samples included 24 samples of seaweed and 11 samples of various species such as sea sponge, sea cucumber, oyster, mussel, shrimp

and clams. All investigated seaweed species belong to the class of brown seaweed (Phaeophyta), 118 except for sample 35 which was nori, a red seaweed. Among the samples of seaweed, ten samples of 119 120 Sargassum sp., five samples of Laminaria sp., four samples of Fucus sp. and two samples of Undaria 121 sp. were obtained. All sea sponge samples belong to the class of *Demospongiae*. Most of the purchased samples were dried at the point of purchase. Samples collected directly from the sea were dried at room 122 temperature or freeze-dried after collection. All dried samples (MT 1 to 25, 31 to 35) were stored at room 123 temperature, and all wet samples (MT 26 to 30) were stored at -20°C. 124

125

Sample preparation 126

For sample preparation a previously established method was used with modifications (Fujii et al., 2014). 127 128 In brief, 300 to 400 mg of each dry sample and 1000 to 1500 mg of each wet sample were twice extracted 129 with 6 mL of a ethyl acetate: MeOH: DCM (1:1:1, v/v/v) mixture. After centrifugation, the combined supernatants were evaporated to dryness and reconstituted in 5 mL of hexane. One milliliter of 1M 130 KOH:EtOH (7:3, v/v) was added. After vortexing and centrifugation, the hexane fraction was transferred 131 to another glass tube. The process was repeated with 5 mL of hexane. The hexane fraction was subject 132 133 of another study and will not be discussed.

1.3 mL of 1M HCl and 5 mL of hexane:diethylether (1:1, v/v) were added to the 1M KOH:EtOH (7:3, v/v) fraction. After vortexing and centrifugation, the hexane:diethylether (1:1, v/v) fraction was transferred to another glass tube and the procedure was repeated. The combined fractions were concentrated to nearly dryness and reconstituted in 50 µL of recovery standard (d18-y-HBCD, 150 pg/µL) and 150 µL of MeOH. Due to the complex matrices, samples were diluted with a ratio of 1 to 5 in methanol prior to injection.

A detailed description of the sample preparation procedure can be found in the supporting information. Procedural blank was prepared using the procedure described above.

Instrumentation and LC-QTOF-MS analysis

6

0

1

2

All measurements were conducted on an Agilent Infinitity 1290 UPLC coupled to an Agilent 6530 144 quadrupole time-of-flight (QTOF) mass spectrometer equipped with an electrospray ionization (ESI) 145 146 source (Agilent Technologies, Santa Clara, USA). The final chromatographic method included a Kinetex 147 Biphenyl column (2.1 mm x 50 mm, 1.7 µm particle size; Phenomenex, Utrecht, Netherlands) equipped with a SecurityGuard™ ULTRA guard column (i.d. 2.1 mm; Phenomenex, Utrecht, Netherlands) with the 148 same stationary phase. The mobile phases consisted of water with 5 mM ammonium acetate (A) and 149 methanol (B). The final optimized chromatographic conditions and ion source parameters are 150 151 summarized in Table S3. The QTOF was operated in negative ionization mode. For sample analysis using target and suspect screening, the Auto MS/MS acquisition mode was applied to allow an automatic 152 selection of precursor ions (max. 4 per cycle). The quadrupole was operated in narrow selection mode 153 154 $(m/z \pm 1.3)$ and fragments were recorded in a mass range of m/z 50 to 1500 applying collision energies 155 of 10, 20 and 40 eV. The monoisotopic mass of the [M-H]⁻ ion of each target compound (Table S1) was included as preferred for fragmentation. Based on the results of target and suspect screening, the sea 156 sponge samples were selected for further investigation using non-target screening. Therefore, these 157 158 samples were injected in full scan MS mode (mass range m/z 100-1700) to enhance cluster detection. Additionally, a selection of the most abundant highly brominated compounds detected within non-target 159 d trom mostwiedzy. 161 162 screening were analyzed in target MS/MS mode to obtain further structural information from their MS/MS spectra. Therefore, the most abundant ion was chosen as the target m/z value, and the selection mode was set to narrow $(m/z \pm 1.3)$.

0

Data analysis

Targeted screening: The available standards (Table S1) were analyzed under the same conditions as the biota samples. The acquired retention times and fragmentation spectra were summarized in a compound database (using MassHunter PCDL Manager, B.08.00; Agilent Technologies) together with the name, molecular formula and monoisotopic mass of each standard. The data were matched against

all biota samples using MassHunter Qualitative Analysis Software (B.07.00). Matching criteria of ± 0.1 169 min for the retention time (RT) and ± 5 ppm for the mass error were applied. 170

171

172 Suspect screening: A suspect list was developed which contained 1) mono-hydroxylated, 2) dihydroxylated, and 3) mono-hydroxylated and mono-methoxylated compounds of BPs, BBs, BDEs, and 173 BDDs. Biphenyls, diphenyl ethers and dibenzodioxins were set with up to a total of 7 bromine and/or 174 175 chlorine atoms, but only up to 3 chlorine atoms. Phenols were set with all possible combinations with 176 maximum 5 bromine and/or chlorine atoms. The suspect list can be found in Table S4 (shortened overview) and Table S5 (full version). After chromatogram deconvolution and alignment, the suspect list 177 was matched against the analyzed samples using the "targeted feature extraction" algorithm of the 178 179 Profinder software (B.08.00; Agilent Technologies, USA). Thereby, matching criteria of ± 10 ppm for the 180 mass error and ± 0.2 min for RT alignment were used.

181

Non-target screening: HaloSeeker 1.0 was used to analyze samples acquired in full scan MS mode. 182 183 This R based open source software allows the identification of halogenated (i.e. chlorinated and/or brominated) compounds based on the characteristic mass differences and relative intensities between 184 ³⁵Cl/³⁷Cl and ⁷⁹Br/⁸¹Br isotopes, respectively (Leon et al., 2019). After automatic raw data conversion 186 into mzXML format by built in msConvert software version 3.0.9810 (Chambers et al., 2012), peak 187 picking was performed applying the xcms package (version 3.2.0; m/z tolerance = 3; peakwidth = 5-60; prefilter step = 3; prefilter level = 10 000; sntresh = 10). This provided a list of extracted signals (i.e., 188 features) which was paired based on the specific mass differences and isotopic patterns of C, Br and Cl 189 atoms (RT tolerance = 5 s; m/z tolerance = 0.5 mDa) (Cariou et al., 2016). The results were displayed in MD plots representing the fractional part of an m/z value using a H/Cl scale instead of the common 1 2 IUPAC scaling (Taguchi et al., 2010). This allows to distinguish between alkane and halogenated series as latter align on horizontal lines while alkane series show a diagonal alignment. To the obtained MD 3 plots the F2+ filter was applied. This filter represents the fourth level of filtering available within the 4

HaloSeeker software. While the F0 filter would display all detected features and the F1 filter retain all 195 paired features, the F2 filtering step applies additional rules retaining only paired features showing ion 196 197 ratios corresponding to halogenated clusters. Lastly, the F2+ filter retains only polyhalogenated clusters. 198 For the obtained series, molecular formula (selected elements: H, C, O, Cl, Br, S; m/z tolerance = 10 mDa; relative abundance tolerance = 20%) were predicted. Series of interest which were not fragmented 199 in Auto MS/MS mode were analyzed in target MS/MS mode to obtain further structural MS/MS 200 201 information.

202

Results and discussion 203

Method optimization and quality control 204

For the optimization of the chromatographic separation, different gradient slopes and four different 205 206 columns were tested: Poroshell 120 EC-C18 columns with two different dimensions (i.e., 3.0 mm x 50 mm and 3.0 mm x 100 mm, both with 2.7 µm particle size; Agilent Technologies), a Kinetex Biphenyl 207 column (2.1 mm x 50 mm, 1.7 µm particle size; Phenomenex, Utrecht, Netherlands) and a Kinetex 208 209 Phenylhexyl column (2.1 mm x 50 mm, 1.7 µm particle size; Phenomenex, Utrecht, Netherlands). The aim of method optimization was to achieve a clear separation and satisfactory peak shapes for the two 210 main compound classes covered within target screening (i.e., BPs and OH-PBDEs) rather than the separation of isomers as latter would require long run times as observed in previous studies (Chi et al., 212 213 2017). Ultimately, the Kinetex Biphenyl column (2.1 mm x 50 mm, 1.7 µm) gave the best results 214 regarding peak shape and width and was chosen for further analyses. The retention times (RT) and observed fragments of all available standards are summarized in Table S6.

Using the Kinetex Biphenyl column (Phenomenex, Utrecht, Netherlands) several gradients of different lengths and slopes were tested aiming to achieve separation of the partially coeluting bromophenols and isomers of tetrabrominated OH-BDEs (i.e., 2'OH-BDE68, 4'OH-BDE49, 6OH-BDE47, 5OH-BDE47). As flatter, longer gradients did not allow complete isomer separation while having a negative

7

8

9

influence on the peak shapes, the gradient shown in Table S3 was chosen as it allowed a clear 220 separation between BPs and OH-PBDEs, satisfying peak shapes and a short run time. 221

222

223 The IS was detectable in all samples with an average retention time of 8.24 min \pm 0.02 min. As the IS which was added prior to sample preparation to indicate potential analyte losses was detectable in all 224 samples, the data of all samples was considered. 225

226

0

MOST WIEDZY

2

3

4

5

227 Target screening

The retention times and acquired fragmentation spectra (Table S6) of all available standards were 228 summarized in a compound database. The database was matched against all analyzed samples. This 229 230 allowed the identification of compounds with confidence level (CL) 1 according to Schymanski et al. 231 (Schymanski et al., 2014). However, this assignment was only made if all identifiers (i.e., the retention time, exact mass, fragmentation spectra and isotopic pattern) unequivocally matched the reference 232 data. Since bromine fragments (m/z 78.9189 and m/z 80.9168) were the most abundant in all 233 234 fragmentation spectra and the detection of less abundant fragments in the samples was limited, the unequivocal identification of isomeric structures based on this information is challenging. This is 235 additionally hampered by the limited chromatographic separation of isomers. Also, several samples ≥236 showed overall low abundant fragmentation spectra and missing ions in the isotopic patterns of target 237 238 compounds. Latter was observed for low abundant compounds and is especially problematic as the characteristic isotopic pattern of brominated compounds provides valuable information for their 239 8240 identification. These observations did not allow an unequivocal assignment of CL 1. In this case, level 241 3 was assigned.

To allow inclusion of information about the observed fragments and their abundances in the report of identification certainty, we propose a subdivision of CL 3 in levels 3A to 3D (Table 1). This subdivision was considered useful as it also allows to report differences in the observed isotopic patterns which is of high relevance for the identification of halogenated analytes. Also, the observed abundances of

fragmentation spectra were considered (3A/3B vs. 3C/3D). Thereby, level 3A or 3B was assigned if the most abundant fragment in the fragmentation spectra showed an intensity above 10³ counts. This threshold was chosen based on empirical observations made within the analysis of samples for target screening. An intensity <10³ counts in fragmentation spectra always correlated with the presence of only bromine fragments leading to limited information on the identity of the compounds and thus a lower identification confidence (i.e., CL 3C/3D).

252

253 <u>Table 1:</u> Subdivision of confidence level 3 (Schymanski et al., 2014) proposed for the identification of NHCs within target and 254 suspect screening. For the subdivision, complete (\checkmark) and incomplete (\times) matches of the isotopic pattern as well as the 255 observed intensities of fragmentation spectra were considered. Within target screening a match of the retention time with the 256 corresponding reference standard was necessary (\checkmark).

Confidence level (CL)	Retention time	Fragmentation	Isotopic pattern			
CL 3A	√	High intensity (≥10 ³ counts)	✓			
CL 3B	√	High intensity (≥10 ³ counts)	×			
CL 3C	1	Low intensity (≤10 ³ counts)	✓			
CL 3D	~	Low intensity (≤10 ³ counts)	×			

258 259 260

261

262

263

4

5

6

7

MOST WIEDZY

257

The results of target screening can be found in Tables S7 (OH-PBDEs in algae samples), S8 (OH-PBDEs in sea sponge samples) and S9 (BPs in all samples). None of the compounds reported there were present in the procedural blanks. From the group of OH-PBDEs and dihydroxylated compounds, 6OH-BDE47 and 2,2-diOH-BB80 were the most frequently detected compounds and were found in eleven and thirteen out of 35 samples, respectively. In invertebrates' samples (MT-25 to MT-30) no compounds were detected within target screening. Therefore, only the results of algae and sea sponge samples are discussed in the following sections.

<u>Algae</u>

6OH-BDE47 was detected in seven alga samples of *Sargassum* sp. and *Laminariaceae* sp. All these samples derived from the Asia/Pacific region. These results are in line with previous studies in which

6OH-BDE47 was one of the most frequently detected compounds in a set of 16 aquatic plant samples 268 collected from Luzon Island, Philippines (Haraguchi et al., 2010). 269

270 2,2-diOH-BB80 was detected in ten alga samples (Sargassum fusiforme, Laminariaceae sp., Undaria 271 pinnatifida, Fucus vesiculosus, Fucus spiralis and Himanthalia elongata). Thereby, Sargassum fusiforme and Laminariaceae sp. derived from the Asia/Pacific region whilst Undaria pinnatifida, Fucus 272 vesiculosus, Fucus spiralis and Himanthalia elongate were collected in the Atlantic. The detection of 273 2,2-diOH-BB80 in samples from the Asia/Pacific region again confirms the findings of Haraguchi et al. 274 275 who reported the presence of 2,2-diOH-BB80 and its dimethoxylated analogue 2,2-diMeO-BB80 in aquatic plant samples from the Philippines (Haraguchi et al., 2010). Additionally, another dihydroxylated 276 compound (2,2-diOH-BDE68) was detected in two samples of edible kombu (Laminariaceae sp. and 277 278 Laminaria ochroleuca). These findings can serve as a first indication of a more ubiquitous presence of 279 dihydroxylated and dimethoxylated analogues of BB80 and BDE68 in marine biota from both the Atlantic and the Asia/Pacific. 280

281

0

MOST WIEDZY

282 From the group of BPs (Table S9), monobromophenol was detected in two alga samples (kelp kombu, Laminaria digitata and bladderwrack, Fucus vesiculosus) which both derived from the Atlantic. However, 283 the identification of the position of bromine was not possible, as isomers of monobromophenol were not 284 available as individual standards. 2,4-Dibromophenol (2,4-DBP) was detected in brown seaweed 285 286 (Indonesia) and kelp kombu (Laminaria digitata, Ireland). 2,4,6-Tribromophenol (2,4,6-TBP) was the most frequently detected BP and was found in 9 alga samples, confirming previous findings. For 287 example, within a study on BPs in a large set of marine algae samples (n = 87), 2,4,6-TBP was reported 288 289 with a detection frequency (DF) of 100% (Whitfield et al., 1999). However, this study also reported a DF of 100% for 2,4-DBP which was not confirmed by our data. This might indicate a limited sensitivity of 0 1 the applied sample preparation or instrumental method for the detection of BPs. Further optimizations 2 of the chromatographic conditions and additional reference standards can further improve the results 3 presented here.

Sea sponge 294

It must be noted that two of the sea sponge samples, namely MT-11 (Lamellodysidea sp.) and MT-31 295 296 (Callyspongia sp.), have previously been screened by Kato et al. for (di)OH-, (di)MeO- and OH-MeO-297 PBDEs using tandem mass spectrometry (MS/MS) with atmospheric pressure chemical ionization (Kato et al., 2012). Additionally, MT-31 was included in a study on OH- and MeO-PBDEs conducted by 298 Haraguchi et al. using GC-MS/MS (Haraguchi et al., 2011). Therefore, our study aimed to serve as a 299 complementary approach to identify new OH-NHCs through the application of LC-HRMS. 300

301 The results of target screening of sea sponge samples are summarized in table S8 (OH-PBDEs and dihydroxylated compounds) and table S9 (BPs). From the group of BPs, only 2,4-DBP (CL1) and 302 tetrabromophenol (CL4) as well as 2,4,6-TBP (CL 3D) were detected in Lamellodysidea sp. (MT-11) 303 304 and Demospongiae sp. (MT-32), respectively.

305 6OH-BDE47 was detected in all investigated sea sponge samples. These observations are in agreement with the results of target screening on algae samples (see above) as well as previous studies 306 which described 6OH-BDE47 as the predominant compound in sea sponges and species from higher 307 308 trophic levels (Haraguchi et al., 2011; Wan et al., 2009). Additionally, Haraguchi et al. had identified 6OH-BDE47 in Callyspongia sp. (MT-31) which was confirmed by our study (Haraguchi et al., 2011). 309 This also applies to the confirmed identification of 2,2'diOH-BDE68. Kato et al. identified a mono-HO ≳310 and di-HO-tetrabrominated-BDE in Lamellodysidea sp. (MT-11). However, due to the unavailability of a 311 312 reference standard, no unequivocal isomer identification was possible (Kato et al., 2012). Our study therefore provides additional identification confidence by reporting 6OH-BDE47 and 2,2'diOH-BDE68 in 313 8314 the Lamellodysidea sp. sample. Furthermore, 4OH-BDE17, 6OH-BDE85, 3OH-BDE154 and 2,2'diOH-BB80 were identified in Lamellodysidea sp. (MT-11; CL 1 for all compounds, except for 6 3OH-BDE154) which have not been described by Kato et al. Since the focus of their study of was mainly 7 placed on diOH- and diMeO-PBDEs, the mono-OH congeners were not reported (Kato et al., 2012). 8 Our study thus provides additional information on OH-PBDEs and a tetrabrominated di-OH-BB present

in sea sponge samples of Lamellodysidea sp. indicating the advantages of the applied comprehensive 319 screening approach. 320

321

322 Within the analysis of the Lamellodysidea sp. sample (MT-11), strong matrix effects were observed which were assumed to hamper the detection and identification of lower abundant NHCs. After an 323 additional dilution step (final dilution factor of 25), an additional tribrominated OH-PBDE (3OH-BDE28; 324 325 CL 3B) was identified which had not been described by Kato et al. (Kato et al., 2012).

326 The detection of 6OH-BDE47 in samples of *Demospongiae* sp. (MT-32 and MT-33; CL 3B and 1, respectively) confirmed the ubiquitous occurrence of this compound. 6OH-BDE85, 2,2'-diOH-BDE68, 327 and 2,2'-diOH-BB80 were detected in both Demospongiae sp. samples with CL 3D, 4 and 3D, 328 329 respectively. Additionally, 3OH-BDE154 was identified in one of the Demospongiae sp. samples (MT-32; CL 3D). Even though these results have to be interpreted carefully since the identified compounds all 330 showed low levels of identification confidence, to our knowledge this is the first report of mono-OH and 331 di-OH BDEs and BBs in sea sponges of Demospongiae sp. Further studies are needed to provide higher 332 333 identification levels as well as quantitative data on the described compounds.

334 335

336

337

338

339

∂340

1

2

3

MOST WIEDZY

0

Suspect screening

For suspect screening, the created suspect list (Table S5) was matched against all sample files acquired in Auto MS/MS mode. To further increase identification certainty, only matches with a matching score ≥75 were considered. This matching score was calculated by the data analysis software (Profinder B.08.00; Agilent Technologies, USA) based on the accuracy of the experimental mass and isotopic pattern in comparison to theoretical values. As a matching score \geq 75 always included a complete match of the isotopic pattern, confidence levels 3A or 3C were assigned depending on the intensities observed in fragmentation spectra. If no fragmentation spectra were acquired, CL 4 was assigned. As no reference MS or MS/MS library data was available, an assignment of CL 2 was not possible for the compounds identified through suspect screening.

0

MOST WIEDZY

366

7

8

9

0

Algae 346

The compounds detected in algae samples and one invertebrate sample (MT-34; sea cucumber) are 347 348 summarized in Table S10. The [M-H]⁻ ion of a hexabrominated diOH-BDE with the molecular formula $C_{12}H_4Br_6O_3$ (denoted as **OH-Br₆-BDE**) was detected in three alga samples. The observed isotopic 349 pattern (in comparison with theoretical data), the proposed structure and the fragmentation spectrum of 350 351 **OH-Br**₆-BDE in sample MT-13 (kombu; *Laminaria ochroleuca*) are found in Figure S1. The high match 352 of the isotopic pattern between experimental and theoretical data confirms the number of bromines. The observed fragments with m/z 248.8548 (proposed formula: [C₆H₃Br₂O]⁻; Δppm -3.2 ppm) and m/z 353 406.6717 (proposed formula: $[C_6HBr_4]$; Δppm -7.1 ppm) indicate two phenolic moieties with two and 354 355 four bromines, respectively. Additionally, within target screening, the fragment with theoretical m/z356 248.8556 ([C₆H₃Br₂O]⁻) was only observed for standards which showed a hydroxy group in ortho- (2,2'diOH-BDE68) or *meta*-position (e.g., 5OH-BDE47, 3'OH-BDE28). For standards with a hydroxy group 357 in para-position, the mentioned fragment was not observed even though a dibrominated phenolic moiety 358 359 was present (e.g., 4'OH-BDE49). Therefore, the hydroxy group is assumed to be in the ortho- or metaposition. In sample MT-35, OH-Br₆-BDE was the most abundant compound whose chromatographic 360 peak showed significant tailing indicating column overload. A further dilution of MT-35 was needed (final 361 dilution factor 25) to achieve satisfying chromatographic results. The observed isotopic pattern and 362 363 fragmentation spectra were similar to sample MT-13 (data not shown) resulting in the same assigned level of identification confidence (Table S10). In sample MT-12, no fragmentation spectra of 364 OH-Br₆-BDE could be acquired resulting in an assignment of CL 4. 365

Dihydroxylated PBDEs, such as 2',6-diOH-BDE68, have been reported in sea sponge samples (Haraguchi et al., 2011). A dihydroxylated hexabrominated BDE whose structure corresponds to the mass spectral data observed in this study has been described as a major constituent in the Dysideidae sponge family (Hanif et al., 2007). However, dihydroxylated hexabrominated BDEs have not been reported in alga samples yet, even though penta- and hexabrominated OH-PBDEs have been described

as predominant congeners in specific samples from low trophic levels (Dahlgren et al., 2016). To our knowledge, our study is the first report of a dihydroxylated hexabrominated BDE in alga samples. Further studies are necessary to give more insights on possible formation pathways and to increase the identification certainty of some of the reported compounds.

Furthermore, a chlorinated phenol (denoted as **CP-1**) with the molecular formula C_6H_5 CIO was detected in samples MT-8 (brown seaweed) and MT-34 (sea cucumber, *Holothuria atra*). The observed RT and mass error can be found in Table S10. Figure S2 shows the match between theoretical and observed isotopic patterns confirming the presence of chlorine. However, due to low abundance of the observed signals no fragmentation spectra were obtained leading to an identification with CL 4.

380

381 Sea sponge

Suspect screening of sponge samples yielded 18 identified compounds (Table 2). All these compounds were present in the *Lamellodysidea* sp. sample (MT-11), while eight compounds were identified in sample MT-31 (*Callyspongia* sp.). Compounds assigned with CL 3A which allowed to propose a possible chemical structure based on the observed fragmentation spectra are summarized in Figure 1.



1

2

3

386

387 Figure 1: Proposed chemical structures of compounds detected in sea sponge samples through suspect screening and 388 assigned with CL 3A. 389 In addition to 2.2'-diOH-BDE68 (see target screening), tri-, penta-, hexa- and heptabrominated diOH-BDEs were identified in Lamellodysidea sp. (MT-11). Whilst there were no fragmentation spectra 390 available for diOH-tribrominated-BDE resulting in CL 4, the fragmentation spectra of penta- (1) and 391 hexabrominated (2) diOH-BDE (Figure S3) showed a characteristic fragment of [C₆H₃Br₂O]⁻ (theoretical 392 393 m/z 248.8556) indicating that both compounds carry a dibrominated phenolic moiety. Accordingly, fragments corresponding to a tri- and tetrabrominated phenolic moiety were detected in the penta- and 394 395 hexabrominated diOH-BDE, respectively. As discussed for the hexabrominated diOH-BDE (OH-Br₆-396 **BDE**) in the Algae section above, the characteristic fragment of $[C_6H_3Br_2O]^-$ indicates an ortho- or meta-397 position of the hydroxy group. The described fragments also confirm that the hydroxy groups are located 398 on different aromatic rings.

The fragmentation spectrum of diOH-hepta-BDE (3) (Figure S4) showed fragments of tetra- and tribrominated phenolic moieties, again indicating the positioning of hydroxy groups on different aromatic rings. The fragmentation spectra of all diOH-BDEs showed fragments corresponding to [Br]⁻ and a loss of HBr confirming the bromination of the parent compounds. The fragmentation spectra of penta-(1) and hexabrominated (2) diOH-BDEs (Figure S5) detected in Callyspongia sp. (MT-31) showed similar

fragments as in sample MT-11 allowing the same conclusions as described above. The same applies 404 fragmentation diOH-pentabrominated-BDE MT-33 405 the spectra of in sample and to 406 diOH-hexabrominated-BDE in samples MT-32 and MT-33, respectively (both Demospongiae). There, 407 similar fragments as described above were observed, whereof similar conclusions about the structures 408 of the detected compounds can be drawn.

Journal Pre-proof

409 Table 2: Summary of suspect screening results of sea sponge samples. For each compound the retention time (RT), mass error and level of identification confidence

410 are reported. n.d. = not detected.

			MT-11		MT-31 sea sponge			MT-33 sea sponge			MT-32			
			sea sponge								sea sponge			
			Lame	llodyside	a sp.	Callyspongia sp.			Demospongiae			Demospongiae		
Compound	Formula	Monoisot.	RT	Mass Conf.	Conf.	nf. RT	Mass	Conf.	RT	Mass	Conf.	RT	Mass	Conf.
		mass	[min]	error	level	[min]	error	level	[min]	error	level	[min]	error	level
				[ppm]			[ppm]			[ppm]			[ppm]	
Br3-diOH-diphenyl ether	C ₁₂ H ₇ Br ₃ O ₃	435.7945	5.52	-0.57	4	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.
Br5-diOH-diphenyl ether	C ₁₂ H ₅ Br ₅ O ₃	591.6156	5.86	0.94	3A	6.29	-1.59	3A	6.32	-0.94	4	6.33	-0.48	3A
			6.74	-3.44	ЗA				6.81	-3.22	4			
Br6-diOH-diphenyl ether	$C_{12}H_4Br_6O_3$	669.5261	6.69	1.81	3A	6.55	0.79	3A	6.69	-2.73	3A	6.72	-4.2	ЗA
Br7-diOH-diphenyl ether	C ₁₂ H ₃ Br ₇ O ₃	747.4366	7.23	-3.06	3A	7.27	1.4	4	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.
Br4-OH-MeO-diphenyl ethers	C ₁₃ H ₈ Br ₄ O ₃	527.7207	7.83	-0.85	3A	7.89	-3.66	4	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.
Br5-OH-MeO-diphenyl ethers	C ₁₃ H ₇ Br ₅ O ₃	605.6312	8.10	-1.15	3A	8.11	-2.63	4	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.
			8.44	-5.72	4	8.45	-4.68	4						
Br6-OH-MeO-diphenyl ethers	C ₁₃ H ₆ Br ₆ O ₃	683.5417	7.08	-0.95	4	7.12	-4.04	4	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.
			8.25	-6.61	4	8.90	-3.28	4						
			8.89	-7.86	4									
Br ₃ -OH-diphenyl ether/	$C_{12}H_5Br_5O_2$	575.6206	7.27	-4.58	3C	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.
Br₅-diOH-biphenyl														
CI-Br₅-diOH-diphenyl ethers	C ₁₂ H ₄ Br ₅ ClO ₃	625.5766	6.36	-1.18	3A	6.48	-3.42	3A	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.
Br ₃ -OH-MeO-biphenyl	$C_{13}H_9Br_3O_2$	433.8153	7.87	0.49	4	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.
Br ₄ -OH-MeO-biphenyl	$C_{13}H_8Br_4O_2$	511.7258	8.48	-3.3	4	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.
Cl-Br ₃ -OH-MeO-biphenyl	$C_{13}H_8Br_3CIO_2$	467.7763	8.29	-4.32	4	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.
Cl ₂ -Br ₂ -OH-MeO-biphenyls	$C_{13}H_8Br_2Cl_2O_2$	423.8268	8.10	-2.36	4	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.
Br5-diOH-dibenzo-dioxins	$C_{12}H_3Br_5O_4$	605.5948	5.41	4.04	4	5.45	-2.22	4	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.
Br ₂ -diOH-phenol	$C_6H_4Br_2O_2$	265.8578	2.87	1.66	3A	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.
Br3-diOH-phenol	C ₆ H ₃ Br ₃ O ₂	343.7683	4.88	-1.31	3A	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.
Br ₃ -OH-MeO-phenol	C7H5Br3O2	357.7840	5.37	-5.62	3A	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.	n.d.
In Lamellodysidea sp. (MT	-11) tetra, nenta	- and heyahro	minato)Fe wa	ra data	ctad T	horoh	nonta	- and b	hovahr	ominat	۵d

In Lamellodysidea sp. (MI-11), tetra-, penta- and hexabrominated OH-MeO-BDEs were detected. Thereby, p

OH-MeO-BDE showed two and three peaks, respectively, indicating the presence of various isomers. Hexabrominated OH-MeO-BDE 412 and one of the isomers of pentabrominated OH-MeO-BDE were assigned with CL 4, as no fragmentation spectra could be acquired. 413 The fragmentation spectra of both tetra- (4) and the first isomer of pentabrominated (5) OH-MeO-BDE (Figure S6) showed fragments 414 415

which correspond to the loss of the methyl group and the loss of [-CH₃Br]. These observations confirm the methylation of the detected

416 compounds. Additionally, both compounds showed a fragment corresponding to [C₆H₂Br₂O₂]⁻ (theoretical m/z 263.8427) indicating that both compounds carry a dibrominated aromatic moiety. For 417 418 pentabrominated OH-MeO-BDE (5), this was further confirmed by the detection of a fragment 419 corresponding to a tribrominated phenolic moiety ($[C_6H_2Br_3O_2]$; theoretical m/z 341.7532). Interestingly, 420 both described fragments still carried two oxygens while this was not observed for non-methylated 421 diOH-BDEs, indicating different fragmentation pathways. Tetra-, penta- and hexabrominated OH-MeO-BDE were also detected in Callyspongia sp. (MT-31). However, due to low abundance and thereof 422 no available fragmentation spectra, CL 4 was assigned, not allowing any assumption about the structure 423 424 of the compounds.

In Lamellodysidea sp. (MT-11), tetra-, penta- and hexabrominated OH-MeO-BDEs were detected. 425 Thereby, penta- and hexabrominated OH-MeO-BDE showed two and three peaks, respectively, 426 427 indicating the presence of various isomers. Hexabrominated OH-MeO-BDE and one of the isomers of 428 pentabrominated OH-MeO-BDE were assigned with CL 4, as no fragmentation spectra could be acquired. The fragmentation spectra of both tetra- (4) and the first isomer of pentabrominated (5) OH-429 MeO-BDE (Figure S6) showed fragments which correspond to the loss of the methyl group and the loss 430 431 of [-CH₃Br]. These observations confirm the methylation of the detected compounds. Additionally, both 432 compounds showed a fragment corresponding to $[C_6H_2Br_2O_2]^-$ (theoretical m/z 263.8427) indicating that both compounds carry a dibrominated aromatic moiety. For pentabrominated OH-MeO-BDE (5), this 433 was further confirmed by the detection of a fragment corresponding to a tribrominated phenolic moiety 434 435 ($[C_6H_2Br_3O_2]$; theoretical m/z 341.7532). Interestingly, both described fragments still carried two 436 oxygens while this was not observed for non-methylated diOH-BDEs, indicating different fragmentation pathways. Tetra-, penta- and hexabrominated OH-MeO-BDE were also detected in Callyspongia sp. 437 (MT-31). However, due to low abundance and thereof no available fragmentation spectra, CL 4 was 438 439 assigned, not allowing any assumption about the structure of the compounds.

iedzy. 1d. 2441

442

443

444

445

²446

7

8

9

0

MOST WIEDZY

440

It has to be noted that OH-MeO-tetra-BDE, diOH-penta-BDE, OH-MeO-penta-BDE and diOH-hexa-BDE have previously been reported in in both *Lamellodysidea* sp. (MT-11) and *Callyspongia* sp. (MT-31) (Kato et al., 2012). Our study has confirmed these findings and provided further information on fragmentation spectra and on the proposed structures of detected compounds. Additionally, our study reports the detection of a OH-MeO-hexabrominated-BDE and a diOH-heptabrominated-BDE in both sponge species. This compound was not detected by the study of Kato et al. (Kato et al., 2012). However, this finding has to be interpreted carefully as the reported compound could only be assigned with CL4.

In addition to the report of 6OH-BDE85 (RT 8.17 min) within target screening, a second compound with the molecular formula $C_{12}H_5Br_5O_2$ (RT 7.27 min) was detected in *Lamellodysidea* sp. (MT-11). The

fragmentation spectra only yielded fragments corresponding to [Br]. These findings do not allow to 451 unequivocally distinguish whether the observed compound is an additional isomer of a pentabrominated 452 453 OH-BDE or a diOH-BB.



Figure 2: Experimental data of a pentabrominated monochlorinated diOH-BDE with the molecular formula C12H4Br5CIO3 detected in sample MT-11, Lamellodysidea sp. (A) and MT-31, Callyspongia sp. (B). The proposed structure and observed fragmentation spectrum at a collision energy of 40 eV are displayed. The most abundant fragments are assigned with the corresponding neutral losses or proposed structures. The most abundant fragments are assigned with the corresponding neutral losses or proposed structures.

A pentabrominated monochlorinated diOH-BDE (6) whose fragmentation spectra are shown in Figure 2 was detected in Lamellodysidea sp. (MT-11) and Callyspongia sp. (MT-31). Fragments which correspond to the loss of HBr and HCl demonstrated the mixed halogenation of the detected compound. The observed fragments with m/z 248.8500 ([C₆H₃Br₂O]⁻) and m/z 360.7177 ([C₆HBr₃CIO]⁻) indicate that the two phenolic moieties carry two and three bromines plus one chlorine, respectively. Again, the

4

characteristic fragment of [C₆H₃Br₂O]⁻ indicates an ortho- or meta-position of the hydroxy group. To our 466 knowledge, this is the first study reporting a mixed halogenated diOH-BDE in sea sponge species. 467

468

0

MOST WIEDZY

469 Four OH-MeO-BBs with the molecular formulae C₁₃H₉Br₃O₂, C₁₃H₈Br₄O₂, C₁₃H₈Br₃ClO₂ and $C_{13}H_8Br_2Cl_2O_2$ were detected in Lamellodysidea sp. (MT-11). Even though no fragmentation spectra 470 were acquired, and CL 4 had to be assigned, it was assumed that the predicted molecular formulae 471 correspond to OH-MeO-BBs and not MeO-BDEs as latter would not be detectable as [M-H]⁻ ions with 472 473 ESI due to the lack of a free hydroxy group. The observed limitations of the acquisition of fragmentation spectra for these compounds are assumed to be caused by their limited ionization or low abundance in 474 the sample. Further optimization of the sample preparation and instrumental methods is needed to 475 476 confirm the postulated findings. The same applies to the detection of a diOH-hexabromo-dibenzo-dioxin 477 (C12HBr5O4) which was detected in both Lamellodysidea sp. (MT-11) and Callyspongia sp. (MT-31). As mentioned above, the level of bromination was confirmed by the high match of the isotopic pattern. 478 479 However, a more confident confirmation of the structure is not possible due to limited mass spectral 480 information.

Last, diOH-dibrominated-BP (7), diOH-tribrominated-BP (8) and OH-MeO-tribrominated-BP (9) were 481 ₹482 identified in Lamellodysidea sp. (MT-11; all CL 3A). The fragmentation spectra of the dihydroxylated 483 compounds showed fragments corresponding to the loss of HBr, while the fragmentation spectrum of 484 OH-MeO-tribrominated-BP confirmed its methylation by showing a characteristic loss of the methyl group. These findings give insights on new groups of compounds (i.e., diOH-BPs and OH-MeO-BPs) 485 which have not been included in the studies conducted by Kato et al. (2012) and Haraguchi et al. (2011). 486 [≥] 487 The high number and variety of confirmed and newly identified compounds in sea sponge samples indicates the high added value of suspect screening analysis by HRMS. Additionally, sea sponge 8 9 samples of Lamellodysidea sp. (MT-11) and Callyspongia sp. (MT-31) were analysed by non-target 0 screening with the aim to identify further NHCs which were not included in the suspect list.

492 Non-target screening results of sea sponge samples

493 For non-target screening, samples of Lamellodysidea sp. (MT-11) and Callyspongia sp. (MT-31) were 494 analysed in full scan MS mode to enhance feature detection using HaloSeeker 1.0 software, as a 495 complementary approach to suspect screening. Since the results of non-target screening in the mass range below m/z 800 mainly yielded compounds which were already reported through target and 496 suspect screening, the analysis was focused on a mass range between m/z 800 and 1500. Additionally, 497 only polyhalogenated compounds which remained after applying the F2+ filtering within HaloSeeker 1.0 498 499 were considered. Molecular formulae were predicted for all detected features. If fragmentation spectra of sufficient quality were available for a feature and fragments could unequivocally be assigned, the 500 most probably candidate was selected from the proposed formulae based on the assigned fragments. 501 502 For the remaining features, molecular formulae which showed the highest matching score are reported. 503 For Lamellodysidea sp. (MT-11), the non-target screening approach yielded 31 compounds (L1 to L31) (Table S11). The corresponding H/CI-scale can be found in Figure S8. Additionally, fragmentation 504 spectra with satisfying intensities could be acquired for 15 compounds using the target MS/MS mode. 505 506 Based on the predicted molecular formulae, compounds L1 to L4 were identified as hepta-, octa- and nonabrominated dihydroxylated diphenoxybenzene. This was confirmed by analysis of the 507 corresponding masses in target MS/MS mode which yielded fragmentation spectra shown in Figure 3. ≥508 509 Based on this information, hepta- ($C_{18}H_7Br_7O_4$, Δppm -2.61 ppm and -3.08 ppm), octa- ($C_{18}H_6Br_8O_4$, 510 Δppm -1.19 ppm) and nonabrominated $(C_{18}H_5Br_9O_4,$ Δppm -7.98 ppm) dihydroxylated diphenoxybenzenes were detected in Lamellodysidea sp. (MT-11). For heptabrominated dihydroxylated 511 8512 diphenoxybenzene, the observed fragments with molecular formulae $[C_6H_3Br_2O_2]^-$ (theoretical m/z²513 266.8485) and $[C_6H_2Br_3O_2]^-$ (theoretical m/z 344.7590) gave evidence about the distribution of bromines 4 between the three aromatic moieties. The same applies to octabrominated dihydroxylated 5 diphenoxybenzene, for which fragments with molecular formulae $[C_6H_3Br_2O_2]^-$ (theoretical m/z 6 266.8485) and $[C_6HBr_4O_2]^-$ (theoretical m/z 424.6675) were observed. This information provided

517 additional confirmation of compound identification and was used to propose the structures given in 518 Figure 3. To our knowledge, this is the first time that OH-PBDPB are reported in the environment.

519 For compounds **L5** to **L7**, the observed fragmentation spectra showed low abundances not providing 520 additional structural information. The mass difference between compounds **L6** and **L7** and their diagonal 521 orientation in the H/Cl-plot indicated that their molecular formulae differed by a methylene group. 522 Molecular formulae with the highest matching score and which were in line with the described mass 523 differences were selected and are reported in Table S11.



Figure 3: Experimental data and proposed chemical structures of hepta- and octabrominated dihydroxylated diphenoxybenzenes (diOH-BDPBs) with the molecular formulae $C_{18}H_7Br_7O_4$ (A) (L1/L2) and $C_{18}H_6Br_8O_4$ (B) (L3) detected in *Lamellodysidea* sp. (sample MT-11). The most abundant fragments are assigned with the corresponding neutral losses or proposed structures.

5

6

7

8

The fragmentation spectra of compounds L8 to L11 are found in Figures S8 and S9, respectively. All 530 fragmentation spectra showed a fragment corresponding to [HSO4]⁻ (theoretical m/z 96.9601) indicating 531 532 that all four compounds carry at least one sulfate group. Additionally, the fragmentation spectra showed 533 a fragment which could be assigned with the formula $[C_{12}H_4Br_5O_3]^-$. This formula and m/z corresponded to pentabrominated diOH-BDE. These findings were in line with the predicted formulae of compounds 534 535 L8 to L11 which all included five bromines. The second highly abundant fragment observed in the fragmentation spectra was assumed to represent the non-brominated part of the molecule and differed 536 by a methylene group when comparing compounds L10 and L11. However, the acquired data did not 537 allow a further characterization of the structure and the proposal of a structural formula. Compound L12 538 showed similar fragmentation as described for compounds L8 to L11 (data not shown) with a fragment 539 540 which was assigned with the formula $[C_{12}H_3Br_6O_3]^2$ (theoretical m/z 674.5127) indicating that compound L12 carries an additional bromine in comparison to compound L11. This was confirmed by the fact that 541 both compounds clustered on the same horizontal line in the H/CI-plot. 542

543 Based on the horizontal clustering of compounds L13 to L16 in the H/Cl-plot, they were assumed to 544 have similar molecular formulae with an increasing number of bromines. Compounds L14, L15 and L16 545 showed only two fragments within their fragmentation spectra, of which one corresponded to [Br]. The second fragment was m/z 528.6638 for L14 and m/z 608.5827 for L15 and L16, respectively. This े546 confirmed that the compounds differed by a bromine but did not allow further assumptions about the 547 548 structures. Therefore, molecular formula which showed the highest matching score and corresponded to the observed increasing degree of bromination were assigned to compounds L13 and L16 (see Table 549 S11). 550

Compounds L19 and L20 differed by two bromine atoms. While no satisfying fragmentation spectrum was obtained for compound L19, the fragmentation spectrum of compound L20 showed two fragments of which one corresponded to $[Br]^{-}$. The second fragment with m/z 530.7074 could not unequivocally be assigned with a formula or structure. Therefore, the same approach as described previously was applied for the selection of predicted molecular formulae.

0

ຂີ່ 551

2

3

4

5

MOST WIEDZY

For the remaining compounds (i.e., L17, L18, L21 - L31) the information received from the predicted 556 formulae or fragmentation spectra were not sufficient to allow a proposal of a molecular structure. 557 558 Therefore, for these compounds the molecular formulae which showed the highest matching scores are 559 reported (Table S11). Further analyses are needed to provide additional information on these compounds. 560

561

Non-target analysis of Callyspongia sp. (MT-31) yielded 20 compounds (C1 to C20) which are 562 summarized in Table S12. The corresponding H/Cl-plot can be found in Figure S10. Fragmentation 563 spectra with satisfying intensities could be acquired for 9 compounds using the target MS/MS mode. 564

Similar to the results observed for Lamellodysidea sp. (MT-11), in Callyspongia sp. (MT-31) compounds 565 C1 to C4 were identified as hepta-, octa- and nonabrominated dihydroxylated diphenoxybenzene. The 566 567 fragmentation spectra obtained for hepta- and octabrominated dihydroxylated diphenoxybenzenes (Figure S11) showed fragments identical to the ones observed for sample MT-11. Therefore, the same 568 assumptions regarding the molecular structures and distribution of bromine as described above can be 569 570 made. Additionally, a heptabrominated monochlorinated dihydroxylated diphenoxybenzene (compound C5) with the molecular formula $C_{18}H_6Br_7CIO_4$ was detected. The given formula was proposed based on 571 the grouping of the described compound on the same horizontal line in the H/Cl-plot as the non-े572 573 chlorinated diOH-diphenoxybenzenes indicating the same degree of halogenation. Additionally, the 574 experimental and theoretical isotopic patterns ($\Delta ppm - 8.15 ppm$; Figure S12) showed a satisfying fit. The obtained fragmentation spectrum of compound C5 is shown in Figure 4. Similar to the results 575 obtained for hepta- and octabrominated diOH-BDPBs, a fragment with the molecular formulae °576 $[C_6H_3Br_2O_2]^-$ (theoretical m/z 266.8485) was observed. A fragment which was assigned with the formula $[C_6HBr_3CIO_2]$ (theoretical m/z 380.7178) confirmed that one of the aromatic moieties is carrying one 8 9 chlorine atom.



Figure 4: Experimental data and proposed chemical structure of a heptabrominated monochlorinated dihydroxylated 581 582 diphenoxybenzene (diOH-BDPBs) with the molecular formula C18H6Br7CIO4 (C5) Callyspongia sp. (MT-31). The most abundant 583 fragments are assigned with the corresponding neutral losses or proposed structures.

Based on the grouping in the H/CI-plot and the observed mass differences, compounds C9 and C10 584 were assumed to differ by one bromine atom. As described for compounds L14 to L16 in Lamellodysidea 585 586 sp. (MT-11), the fragmentation spectra obtained for compounds **C9** and **C10** in *Callyspongia* sp. (MT-31) again showed only two fragments (namely, bromine and m/z 528.6691 for compound **C9** and m/z for 587 compound C10) not providing further structural information. 588

Compound **C8** was assigned with molecular formula $C_{24}H_8Br_{10}O_5$ which showed a satisfying match 589 between theoretical and experimental isotopic patterns (Δppm -4.92 ppm; Figure S13). This formula 590 591 was assumed to correspond to a decabrominated compound consisting of four aromatic rings which are 592 linked by ether bonds and of which two carry a hydroxy group. This indicates another novel compound consisting of several aromatic rings detected in *Callyspongia* sp. However, there was no fragmentation 593 spectra available to confirm the proposed structure.

For the remaining compounds available fragmentation spectra did not provide sufficient information for the assignment of probable structural formulae. Therefore, all remaining compounds (i.e., C5 - C8, C11

- C20) were assigned with molecular formulae which showed the highest matching scores.

In conclusion, nine compounds were detected in both sponge samples showing similar isotopic patterns and retention times. Accordingly, 22 compounds which were unique to Lamellodysidea sp. (sample MT-11) and 11 compounds which were unique to *Callyspongia* sp. (MT 31) are reported.

7

8

9

0

601 Conclusions

0

619

621

622

4

5

6

MOST WIEDZY

This study introduced a comprehensive combined screening approach applying target, suspect and 602 603 non-target screening for the identification of new hydroxylated NHCs in biota samples. The use of liquid 604 chromatography high-resolution mass spectrometry allowed the detection of a high variety of NHCs from different classes. 605

Target screening allowed the identification of seven compounds with 6OH-BDE47 being the most 606 607 frequently detected compound. Additionally, the reported results showed the identification of two 608 dihydroxylated compounds (2,2-diOH-BB80 and 2,2-diOH-BDE68) in matrices in which they have not been reported previously. The confirmation of previous results from quantitative studies showed that the 609 approach of target screening gives a reliable overview of NHCs in biota samples with high confidence 610 611 levels of identification.

612 Suspect screening yielded two compounds detected in alga samples (hexabrominated diOH-diphenyl ether and monochlorinated phenol). A high number of compounds was detected in sponge samples (17 613 and 8 compounds identified in sea sponge samples of Lamellodysidea sp. and Callyspongia sp., 614 615 respectively) indicating a high variety of NHCs occurring in this species. Four of the identified 616 compounds have been described in previous studies. Thus, the presented work introduces a high ≧617 number of newly identified NHCs in sea sponge samples including heptabrominated diOH-BDE, monochlorinated pentabrominated diOH-BDE, hexabrominated OH-MeO-BDE and others. 618

Non-target screening allowed the identification of OH-PBDBPs, such as hepta-, octa- and nonabrominated diOH-BDBPs, in Lamellodysidea sp. and Callyspongia sp. samples. To our knowledge, 620 this is the first study reporting these compounds in the environment. Non-target screening yielded additional and 16 compounds in Lamellodysidea sp. and Callyspongia sp. samples, respectively, which could tentatively be identified through the assignment of predicted molecular formulae. 3

This study provides a comprehensive screening approach for polyhalogenated NHCs in biota samples which provided additional information on the occurrence and distribution of NHCs in alga and sea sponge species and can serve as a valuable tool for future screening studies.

Acknowledgments 628

629 LB acknowledges funding through a Research Foundation Flanders (FWO) fellowship (11G1821N). YF 630 acknowledges a research fellowship (Japan Society for the Promotion of Science (JSPS) overseas research fellowship; no. 201860307). MS acknowledges support by the Iwanowska Programme (Polish 631 National Agency for Academic Exchange (NAWA), no. PPN/IWA/2018/1/00089/U/00001). Further 632 financial support was provided by the Ministry of Health, Labour and Welfare of Japan (Y.F.; no. H29-633 food-young researchers-008), the JSPS (Y.F; no. 20K12188 and 21K12262) and by the University of 634

Antwerp (UA). 635

The authors acknowledge Dr. Ronan Cariou for his support and provided input on the HaloSeeker 1.0 636

637 software. Yunita Puspitasari is acknowledged for providing the sea cucumber sample.

638

References 639

- Agarwal, V., El Gamal, A.A., Yamanaka, K., Poth, D., Kersten, R.D., Schorn, M., Allen, E.E., Moore, 640 641 B.S., 2014. Biosynthesis of polybrominated aromatic organic compounds by marine bacteria. Nat Chem Biol 10, 640-647. 642
- 643 Agarwal, V., Li, J., Rahman, I., Borgen, M., Aluwihare, L.I., Biggs, J.S., Paul, V.J., Moore, B.S., 2015. Complexity of naturally produced polybrominated diphenyl ethers revealed via mass spectrometry. 645 Environ Sci Technol 49, 1339-1346.
 - Bidleman, T.F., Andersson, A., Brugel, S., Ericson, L., Haglund, P., Kupryianchyk, D., Lau, D.C.P., Liljelind, P., Lundin, L., Tysklind, A., Tysklind, M., 2019a. Bromoanisoles and methoxylated bromodiphenyl ethers in macroalgae from Nordic coastal regions. Environ Sci Process Impacts 21, 881-892.
 - Bidleman, T.F., Andersson, A., Jantunen, L.M., Kucklick, J.R., Kylin, H., Letcher, R.J., Tysklind, M., Wong, F., 2019b. A review of halogenated natural products in Arctic, Subarctic and Nordic ecosystems. Emerging Contaminants 5, 89-115. 2
 - Cariou, R., Méndez-Fernandez, P., Hutinet, S., Guitton, Y., Caurant, F., Le Bizec, B., Spitz, J., Vetter, W., Dervilly, G., 2020. Nontargeted LC/ESI-HRMS Detection of Polyhalogenated Compounds in Marine Mammals Stranded on French Atlantic Coasts. ACS ES&T Water 1, 309–318.

3

- Cariou, R., Omer, E., Leon, A., Dervilly-Pinel, G., Le Bizec, B., 2016. Screening halogenated 656 environmental contaminants in biota based on isotopic pattern and mass defect provided by high 657 resolution mass spectrometry profiling. Anal Chim Acta 936, 130-138. 658
- Chambers, M.C., Maclean, B., Burke, R., Amodei, D., Ruderman, D.L., Neumann, S., Gatto, L., Fischer, 659 660 B., Pratt, B., Egertson, J., Hoff, K., Kessner, D., Tasman, N., Shulman, N., Frewen, B., Baker, T.A.,
- 661 Brusniak, M.Y., Paulse, C., Creasy, D., Flashner, L., Kani, K., Moulding, C., Seymour, S.L., Nuwaysir,
- 662 L.M., Lefebvre, B., Kuhlmann, F., Roark, J., Rainer, P., Detlev, S., Hemenway, T., Huhmer, A.,
- 663 Langridge, J., Connolly, B., Chadick, T., Holly, K., Eckels, J., Deutsch, E.W., Moritz, R.L., Katz, J.E.,
- Agus, D.B., MacCoss, M., Tabb, D.L., Mallick, P., 2012. A cross-platform toolkit for mass 664 665 spectrometry and proteomics. Nat Biotechnol 30, 918-920.
- Chi, X., Liu, J., Yu, M., Xie, Z., Jiang, G., 2017. Analysis of bromophenols in various aqueous samples 666 using solid phase extraction followed by HPLC-MS/MS. Talanta 164, 57-63. 667
- Choo, G., Lee, I.S., Oh, J.E., 2019. Species and habitat-dependent accumulation and biomagnification 668 of brominated flame retardants and PBDE metabolites. J Hazard Mater 371, 175-182. 669
- Dahlgren, E., Lindqvist, D., Dahlgren, H., Asplund, L., Lehtila, K., 2016. Trophic transfer of naturally 670 671 produced brominated aromatic compounds in a Baltic Sea food chain. Chemosphere 144, 1597-1604. 672
- Dingemans, M.M., de Groot, A., van Kleef, R.G., Bergman, A., van den Berg, M., Vijverberg, H.P., 673 Westerink, R.H., 2008. Hydroxylation increases the neurotoxic potential of BDE-47 to affect 674 exocytosis and calcium homeostasis in PC12 cells. Environ Health Perspect 116, 637-643. 675
- 676 Fujii, Y., Nishimura, E., Kato, Y., Harada, K.H., Koizumi, A., Haraguchi, K., 2014. Dietary exposure to 677 phenolic and methoxylated organohalogen contaminants in relation to their concentrations in breast milk and serum in Japan. Environ Int 63, 19-25. 678
 - Gribble, G.W., 2003. The diversity of naturally produced organohalogens. Chemosphere 52, 289-297.
 - Hamers, T., Kamstra, J.H., Sonneveld, E., Murk, A.J., Visser, T.J., Van Velzen, M.J., Brouwer, A., Bergman, A., 2008. Biotransformation of brominated flame retardants into potentially endocrinedisrupting metabolites, with special attention to 2,2',4,4'-tetrabromodiphenyl ether (BDE-47). Mol Nutr Food Res 52, 284-298.
 - Hanif, N., Tanaka, J., Setiawan, A., Trianto, A., de Voogd, N.J., Murni, A., Tanaka, C., Higa, T., 2007. Polybrominated diphenyl ethers from the Indonesian sponge Lamellodysidea herbacea. J Nat Prod 70, 432-435.
 - Haraguchi, K., Kato, Y., Ohta, C., Koga, N., Endo, T., 2011. Marine sponge: a potential source for methoxylated polybrominated diphenyl ethers in the Asia-Pacific food web. J Agric Food Chem 59, 13102-13109.

6 7

8

- Haraguchi, K., Kotaki, Y., Relox, J.R., Jr., Romero, M.L., Terada, R., 2010. Monitoring of naturally 690 produced brominated phenoxyphenols and phenoxyanisoles in aquatic plants from the Philippines. J 691 Agric Food Chem 58, 12385-12391. 692
- Kato, Y., Okada, S., Atobe, K., Endo, T., Haraguchi, K., 2012. Selective determination of mono- and 693 694 dihydroxylated analogs of polybrominated diphenyl ethers in marine sponges by liquid-695 chromatography tandem mass spectrometry. Anal Bioanal Chem 404, 197-206.
- 696 Kato, Y., Okada, S., Atobe, K., Endo, T., Matsubara, F., Oguma, T., Haraguchi, K., 2009. Simultaneous 697 determination by APCI-LC/MS/MS of hydroxylated and methoxylated polybrominated diphenyl ethers found in marine biota. Anal Chem 81, 5942-5948. 698
- 699 Kim, U.J., Jo, H., Lee, I.S., Joo, G.J., Oh, J.E., 2015. Investigation of bioaccumulation and biotransformation of polybrominated diphenyl ethers, hydroxylated and methoxylated derivatives in 700 varying trophic level freshwater fishes. Chemosphere 137, 108-114. 701
- 702 Lacorte, S., Ikonomou, M.G., Fischer, M., 2010. A comprehensive gas chromatography coupled to high 703 resolution mass spectrometry based method for the determination of polybrominated diphenyl ethers 704 and their hydroxylated and methoxylated metabolites in environmental samples. J Chromatogr A 705 1217, 337-347.
- Leon, A., Cariou, R., Hutinet, S., Hurel, J., Guitton, Y., Tixier, C., Munschy, C., Antignac, J.P., Dervilly-706 Pinel, G., Le Bizec, B., 2019. HaloSeeker 1.0: A User-Friendly Software to Highlight Halogenated 707 708 Chemicals in Nontargeted High-Resolution Mass Spectrometry Data Sets. Anal Chem 91, 3500-3507. 709
- 710 Li, F., Xie, Q., Li, X., Li, N., Chi, P., Chen, J., Wang, Z., Hao, C., 2010. Hormone activity of hydroxylated polybrominated diphenyl ethers on human thyroid receptor-beta: in vitro and in silico investigations. ⁰712 Environ Health Perspect 118, 602-606.
- 713 Linares, V., Belles, M., Domingo, J.L., 2015. Human exposure to PBDE and critical evaluation of health 714 hazards. Arch Toxicol 89, 335-356.
 - Liu, M., Hansen, P.E., Lin, X., 2011. Bromophenols in marine algae and their bioactivities. Mar Drugs 9, 1273-1292.
 - Losada, S., Roach, A., Roosens, L., Santos, F.J., Galceran, M.T., Vetter, W., Neels, H., Covaci, A., 2009. Biomagnification of anthropogenic and naturally-produced organobrominated compounds in a marine food web from Sydney Harbour, Australia. Environ Int 35, 1142-1149. 9
 - Lupton, S.J., McGarrigle, B.P., Olson, J.R., Wood, T.D., Aga, D.S., 2010. Analysis of hydroxylated polybrominated diphenyl ether metabolites by liquid chromatography/atmospheric pressure chemical ionization tandem mass spectrometry. Rapid Commun Mass Spectrom 24, 2227-2235.

1

- 723 Malmvarn, A., Marsh, G., Kautsky, L., Athanasiadou, M., Bergman, A., Asplund, L., 2005. Hydroxylated and methoxylated brominated diphenyl ethers in the red algae Ceramium tenuicorne and blue 724 mussels from the Baltic Sea. Environ Sci Technol 39, 2990-2997. 725
- Malmvarn, A., Zebuhr, Y., Kautsky, L., Bergman, K., Asplund, L., 2008. Hydroxylated and methoxylated 726 727 polybrominated diphenyl ethers and polybrominated dibenzo-p-dioxins in red alga and cyanobacteria 728 living in the Baltic Sea. Chemosphere 72, 910-916.
- 729 Schymanski, E.L., Jeon, J., Gulde, R., Fenner, K., Ruff, M., Singer, H.P., Hollender, J., 2014. Identifying 730 small molecules via high resolution mass spectrometry: communicating confidence. Environ Sci Technol 48, 2097-2098. 731
- Su, G., Yu, H., Lam, M.H., Giesy, J.P., Zhang, X., 2014. Mechanisms of toxicity of hydroxylated 732 polybrominated diphenyl ethers (HO-PBDEs) determined by toxicogenomic analysis with a live cell 733 array coupled with mutagenesis in Escherichia coli. Environ Sci Technol 48, 5929-5937. 734
- Sun, H., Li, Y., Hao, Y., Zhu, Y., Yang, R., Wang, P., Zhang, Q., Jiang, G., 2020. Bioaccumulation and 735 Trophic Transfer of Polybrominated Diphenyl Ethers and Their Hydroxylated and Methoxylated 736 Analogues in Polar Marine Food Webs. Environ Sci Technol 54, 15086-15096. 737
- 738 Taguchi, V.Y., Nieckarz, R.J., Clement, R.E., Krolik, S., Williams, R., 2010. Dioxin analysis by gas chromatography-Fourier transform ion cyclotron resonance mass spectrometry (GC-FTICRMS). J 739 Am Soc Mass Spectrom 21, 1918-1921. 740
- 741 Thomsen, C., Lundanes, E., Becher, G., 2001. Brominated flame retardants in plasma samples from 742 three different occupational groups in Norway. J Environ Monit 3, 366-370.
- 743 Van Boxtel, A.L., Kamstra, J.H., Cenijn, P.H., Pieterse, B., Wagner, M.J., Antink, M., 2007. Microarray analysis reveals a mechanism of phenolic polybrominated diphenylether toxicity in zebrafish. Environ Sci Technol 42, 1773-1779.
 - Wan, Y., Wiseman, S., Chang, H., Zhang, X., Jones, P.D., Hecker, M., Kannan, K., Tanabe, S., Hu, J., Lam, M.H., Giesy, J.P., 2009. Origin of hydroxylated brominated diphenyl ethers: natural compounds or man-made flame retardants? Environ Sci Technol 43, 7536-7542.
 - Wang, S., Wu, T., Huang, H., Ping, H., Lu, A., Zhang, S., 2011. Analysis of hydroxylated polybrominated diphenyl ethers in plant samples using ultra performance liquid chromatography-mass spectrometry. Science China Chemistry 54, 1782-1788.
 - Weijs, L., Losada, S., Das, K., Roosens, L., Reijnders, P.J., Santos, J.F., Neels, H., Blust, R., Covaci, A., 2009. Biomagnification of naturally-produced methoxylated polybrominated diphenyl ethers (MeO-PBDEs) in harbour seals and harbour porpoises from the southern North Sea. Environ Int 35, 893-899.

3

Whitfield, F.B., Helidoniotis, F., Shaw, K.J., Svoronos, D., 1999. Distribution of bromophenols in species
of marine algae from eastern Australia. J Agric Food Chem 47, 2367-2373.

758

Journal Prevence

Highlights:

- HRMS and suspect screening used to identify naturally occurring halogenated compounds
- First report of 2,2-diOH-BB80 in alga samples derived from the Atlantic
- Newly identified hydroxylated BDEs in sea sponge
- First detection of polybrominated dihydroxylated diphenoxybenzenes (diOH-PBDPBs)

ournal pre-proof

Declaration of interests

 \boxtimes The authors declare that they have no known competing financial interests or personal relationships that could have appeared to influence the work reported in this paper.

□The authors declare the following financial interests/personal relationships which may be considered as potential competing interests: